

Attachment 8 Acceptable Toxicity Studies for the Selection of Fish Toxicity Reference Values

Table 1. Acceptable toxicity studies for selection of fish dietary toxicity reference values

| CHEMICAL | CHEMICAL FORM                     | TEST SPECIES  | LIFE STAGE | NOAEL (mg/kg dw <sup>a</sup> ) | LOAEL (mg/kg dw <sup>a</sup> ) | EXPOSURE ROUTE | EXPOSURE DURATION | ENDPOINT         | ENDPOINT EFFECT                                      | SOURCE                                | NOTES   |
|----------|-----------------------------------|---------------|------------|--------------------------------|--------------------------------|----------------|-------------------|------------------|--|---------------------------------------|---|
| Arsenic  | sodium arsenite                   | rainbow trout | juvenile   | 20 <sup>b</sup>                | 30 <sup>b</sup>                | food           | 6 wks             | growth           | body weight  | Oladimeji et al. (1984)               | concentrations in figure and text do not agree in study: 20 mg/kg is mentioned both as an effect level and a NOAEL in the text; however, it is shown in the figure to be not significant; 20 mg/kg = assumed NOAEL; food consumption not measured; fish fed prepared diet treated with chemical salts; only the nominal concentration was reported  |
| Arsenic  | disodium arsenate heptahydrate    | rainbow trout | juvenile   | 8                              | 44                             | food           | 16 wks            | growth           | body weight  | Cockell et al. (1991)                 | feed refusal accompanied effects; fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate heptahydrate    | rainbow trout | juvenile   |                                | 49                             | food           | 24 wks            | growth           | body weight  | Cockell et al. (1991)                 | feed refusal accompanied effects; effects attributed to feed refusal; fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate heptahydrate    | rainbow trout | juvenile   |                                | 55                             | food           | 8 days            | growth           | body weight  | Cockell et al. (1992)                 | feed refusal accompanied effects; fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate                 | rainbow trout | juvenile   |                                | 58                             | food           | 12 days           | growth           | body weight  | Cockell and Bettger (1993)            | feed refusal accompanied effects; fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate heptahydrate    | rainbow trout | juvenile   | 32                             | 60                             | food           | 12 days           | growth           | body weight  | Cockell et al. (1992)                 | feed refusal accompanied effects; fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate heptahydrate    | rainbow trout | juvenile   | 33                             | 65                             | food           | 24 wks            | growth           | body weight  | Cockell et al. (1991)                 | feed consumption did not differ from controls; body weight gain reduced at 12 weeks in fish fed 33 mg/kg arsenic but not at 24 weeks (body weight was recovered); fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate                 | rainbow trout | juvenile   |                                | 137                            | food           | 8 days            | growth           | body weight  | Cockell and Hilton (1988)             | greater than 10% mortality at LOAEL; feed refusal accompanied effects; study also reported effects of organic arsenic which was not toxic at high levels; fish fed prepared diet treated with chemical salts  |
| Arsenic  | arsenic trioxide                  | rainbow trout | juvenile   |                                | 180                            | food           | 8 days            | growth           | body weight  | Cockell and Hilton (1988)             | feed refusal accompanied effects; greater than 10% mortality at LOAEL; study also reported effects of organic arsenic which was not toxic at high levels; fish fed prepared diet treated with chemical salts  |
| Arsenic  | disodium arsenate heptahydrate    | striped bass  | juvenile   | 52.3                           | 188.8                          | food           | 6 days            | growth           | body weight  | Blazer et al. (1997)                  | feed refusal accompanied effects; fish fed prepared diet treated with chemical salts  |
| Cadmium  | Cd(NO <sub>3</sub> ) <sub>2</sub> | rockfish      | juvenile   | 0.1 <sup>c</sup>               | 0.5                            | food           | 60 days           | growth           | body weight and length growth rate; condition factor | Kim et al. (2004); Kang et al. (2005) | body weight and body length growth rate were significantly affected in both studies at 25 and 125 mg/kg dw; significance of effect on body weight and body length growth rate at 0.5 and 5 mg/kg dw was not consistent between the two studies; condition factor was reduced at all diet concentrations (0.5, 5, 25, and 125 mg/kg), but was not significant at 25 mg/kg; a significant inverse relationship was observed between weight gain and the exposure concentration of dietary Cd at 25 and 125 mg/kg dw; fish fed prepared diet treated with chemical salts |
| Cadmium  | cadmium chloride                  | rainbow trout | fry        | 55 <sup>b</sup>                |                                | primarily food | 60 days           | survival, growth | 100% survival, body weight and length                | Mount et al. (1994)                   | fish exposed to copper, cadmium, lead, and zinc in water at 23.0, 0.97, 3.32, and 46.3 µg/L in addition to dietary exposure; only no-effect reported; fish fed <i>Artemia</i> exposed to cadmium chloride in water; dietary concentrations corrected for a theoretical 20% loss resulting from depuration of copper from <i>Artemia</i> .   |
| Cadmium  | Cd(NO <sub>3</sub> ) <sub>2</sub> | rockfish      | juvenile   | 125                            |                                | food           | 60 days           | survival         |  | Kim et al. (2004)                     | fish fed prepared diet treated with chemical salts  |
| Cadmium  |                                   | guppy         |            | 171 <sup>b</sup>               |                                | food           | 10-30 days        | growth           |  | Hatakeyama and Yasuno (1982)          | fed exposed <i>Monia macrocopa</i> , significant growth effect at day 10 disappeared at day 20 in fish fed 126 and 171 µg/g dw; fish fed prepared diet treated with chemical salts  |
| Cadmium  | cadmium chloride                  | guppy         | adult      | 210 <sup>b</sup>               |                                | food           | 2 mo              | reproduction     | # live fry, fry mortality, premature embryos         | Hatakeyama and Yasuno (1987)          | fed exposed <i>Chironomus yoshimatsui</i> (midge larvae), no reproductive effect after 2 mo   |

Table 1, continued

| CHEMICAL | CHEMICAL FORM               | TEST SPECIES    | LIFE STAGE | NOAEL<br>(mg/kg dw <sup>a</sup> ) | LOAEL<br>(mg/kg dw <sup>a</sup> ) | EXPOSURE<br>ROUTE | EXPOSURE<br>DURATION | ENDPOINT         | ENDPOINT EFFECT                          | SOURCE                       | NOTES  |
|----------|-----------------------------|-----------------|------------|-----------------------------------|-----------------------------------|-------------------|----------------------|------------------|--|------------------------------|--|
| Cadmium  |                             | Atlantic salmon | juvenile   | 250                               |                                   | food              | 4 months             | growth           | growth rate (body weight)                | Lundebye et al. (1999)       | fish fed prepared diet treated with chemical salts   |
| Cadmium  | cadmium chloride            | guppy           | 2 mo       | 274 <sup>b</sup>                  |                                   | food              | 30 days              | growth           | body weight                              | Hatakeyama and Yasuno (1987) | fed exposed <i>Chironomus yoshimatsui</i> (midge larvae), no growth effect by day 30   |
| Cadmium  |                             | rainbow trout   | juvenile   | 294                               |                                   | food              | 15-30 days           | growth, survival | specific growth rate, survival           | Baldisserotto et al. (2005)  | fish fed prepared diet treated with chemical salts   |
| Cadmium  |                             | rainbow trout   | juvenile   | 471                               |                                   | food              | 28 days              | survival, growth | growth rate, survival                    | Franklin et al. (2005)       | fish fed prepared diet treated with chemical salts   |
| Cadmium  | cadmium chloride            | guppy           | 30 day old | 500 <sup>b</sup>                  | 800 <sup>b</sup>                  | food              | 7 mo                 | reproduction     | cumulative number of fry produced        | Hatakeyama and Yasuno (1987) | fed exposed <i>Chironomus yoshimatsui</i> (midge larvae); cumulative number of fry decreased to about 60% of the control in fish exposed to 80 and 160 µg/L; no discussion of statistics; LOAEL and NOAEL are estimated using figure, no replication   |
| Cadmium  | cadmium chloride            | guppy           | 30 day old |                                   | 1,250 <sup>b</sup>                | food              | 7 mo                 | growth, survival | female growth, female survival           | Hatakeyama and Yasuno (1987) | fed exposed <i>Chironomus yoshimatsui</i> (midge larvae); female body weight decreased - 68% of control at the 48th day in fish exposed to 160 µg/L; 6 of 7 females died that were exposed to 160 µg/L; no discussion of statistics- unclear on growth effects at the other doses; LOAEL is estimated using figure; no effect on male growth, no replication |
| Cadmium  | cadmium nitrate             | rainbow trout   | juvenile   | 786                               | 1,395                             | food              | 30 days              | survival         |  | Szebedinsky et al. (2001)    | 57% survival observed in fish fed 1,395 mg/kg; in a separate experiment reported in this study, 92% survival was reported for juvenile rainbow trout exposed to dietary cadmium concentrations of 1,419 mg/kg dw over a 36-day exposure period; survival data were not statistically analyzed in either experiment   |
| Cadmium  | cadmium nitrate             | rainbow trout   | juvenile   | 1,395                             | 2,265                             | food              | 30 days              | growth           | specific growth rate (weight)            | Szebedinsky et al. (2001)    |  |
| Cadmium  | cadmium sulfate             | rainbow trout   |            |                                   | 10,000                            | food              | 28 days              | survival         |  | Handy (1993)                 | 39% (14 of 36) mortality between day 3 and 23 of exposure period; fish fed prepared diet treated with chemical salts   |
| Chromium | chromium (III)              | grey mullet     | 2 yrs old  | 9.42                              |                                   | food and sediment | 8 wks                | growth           |  | Walsh et al. (1994)          | significant increase in growth of the chromium-exposed fish (not an adverse effect); gray mullet consume large amounts of sediment during feeding and fish were exposed to chromium in both diet and sediments (46mg/kg) simultaneously  |
| Copper   | copper                      | channel catfish | fingerling | 8                                 | 16                                | food              | 16 wks               | growth           | body weight                              | Murai et al. (1981)          | significant effects at 4 wks in 16- and 32-mg/kg treatments; growth gain per feed consumed was significantly lower for fish fed 8 mg/kg; fish fed prepared diet treated with chemical salts  |
| Copper   | copper sulfate pentahydrate | channel catfish | fingerling | 40                                |                                   | food              | 13 wks               | growth           | body weight                              | Gatlin and Wilson (1986)     | fish fed prepared diet treated with chemical salts   |
| Copper   | copper sulfate              | rockfish        | juvenile   | 50                                | 100                               | food              | 60 days              | growth           | body weight (growth rate)                | Kang et al. (2005)           | fish fed prepared diet treated with chemical salts   |
| Copper   | copper sulfate pentahydrate | Atlantic salmon | parr       | 98                                |                                   | food              | 12 wks               | growth, survival | survival, condition factor, body weight, | Lorentzen et al. (1998)      |  |
| Copper   | copper sulfate              | rainbow trout   |            | 200                               |                                   | food              | 32 days              | survival         |  | Handy (1992)                 | only no-effect level concentration reported; fish fed prepared diet treated with chemical salts.   |
| Copper   | copper sulfate pentahydrate | rainbow trout   |            | 664 <sup>b</sup>                  |                                   | food              | 24 wks               | growth           |  | Lanno et al. (1985b)         | significant effects at 16 wks; at 24 wks, growth effects no longer apparent; only no-effect level concentration reported; fish fed prepared diet treated with chemical salts   |
| Copper   | copper sulfate              | rainbow trout   |            | 684                               |                                   | food              | 42 days              | growth           |  | Miller et al. (1993)         | only no-effect level concentration reported; fish fed prepared diet treated with chemical salts  |
| Copper   | copper sulfate pentahydrate | Atlantic salmon | parr       | 691.3 <sup>b</sup>                |                                   | food              | 4 wks                | growth           | length, weight, or condition factor      | Berntssen et al. (1999b)     | only no-effect level concentration reported; fish fed prepared diet treated with chemical salts  |
| Copper   | copper sulfate pentahydrate | Atlantic salmon | fry        | 500                               | 700                               | food              | 3 mos                | growth           |  | Lundebye et al. (1999)       | fish fed prepared diet treated with chemical salts   |

Table 1, continued

| CHEMICAL       | CHEMICAL FORM                | TEST SPECIES      | LIFE STAGE           | NOAEL<br>(mg/kg dw <sup>a</sup> ) | LOAEL<br>(mg/kg dw <sup>a</sup> ) | EXPOSURE<br>ROUTE | EXPOSURE<br>DURATION   | ENDPOINT         | ENDPOINT EFFECT                       | SOURCE                    | NOTES   |
|----------------|------------------------------|-------------------|----------------------|-----------------------------------|-----------------------------------|-------------------|------------------------|------------------|---------------------------------------|---------------------------|---|
| Copper         | copper sulfate pentahydrate  | rainbow trout     |                      | 287 <sup>b</sup>                  | 730 <sup>b</sup>                  | food              | 8 wks                  | growth           |                                       | Lanno et al. (1985b)      | feed refusal and 19% mortality was observed at 1,585 mg/kg-diet; fish fed prepared diet treated with chemical salts   |
| Copper         | copper sulfate pentahydrate  | rainbow trout     |                      | 730 <sup>b</sup>                  |                                   | food              | 8 wks                  | survival         |                                       | Lanno et al. (1985b)      | study showed potential effects at higher concentrations- confounded by food avoidance; feed refusal and 19% mortality were observed at 1,585-mg/kg diet; only no-effect level concentration reported; fish fed prepared diet treated with chemical salts  |
| Copper         | copper sulphate pentahydrate | rainbow trout     |                      |                                   | 796 <sup>b</sup>                  | food              | 16 wks                 | growth           |                                       | Lanno et al. (1985a)      | fish were fed on control diet for 12 d after exposure period; fish fed prepared diet treated with chemical salts  |
| Copper         | copper chloride              | rainbow trout     | fry                  | 352 <sup>b</sup>                  |                                   | primarily food    | 60 days                | survival         | survival                              | Mount et al. (1994)       | fish exposed to copper, cadmium, lead, and zinc in water at 23.0, 0.97, 3.32, and 46.3 µg/L in addition to dietary exposure; mortality observed in fish fed <i>Artemia</i> with Cu concentrations of 830 mg/kg was attributed to elevated copper concentrations in the water; fish fed live <i>Artemia</i> exposed to copper chloride in water; dietary concentrations corrected for a theoretical 20% loss resulting from depuration of copper from <i>Artemia</i> |
| Copper         | copper sulfate               | Atlantic salmon   |                      | 638 <sup>b</sup>                  | 868 <sup>b</sup>                  | food              | 3 mo                   | growth           |                                       | Berntssen et al. (1999a)  | fish fed prepared diet treated with chemical salts  |
| Copper         | copper chloride              | rainbow trout     | fry                  | 800 <sup>b</sup>                  |                                   | primarily food    | 60 days                | growth           | body weight and length                | Mount et al. (1994)       | fish exposed to copper, cadmium, lead, and zinc in water at 23.0, 0.97, 3.32, and 46.3 µg/L in addition to dietary exposure; only no-effect level concentration reported; fish fed live <i>Artemia</i> exposed to copper chloride in water; dietary concentrations corrected for a theoretical 20% loss resulting from depuration of copper from <i>Artemia</i>   |
| Copper         | copper sulfate pentahydrate  | rainbow trout     |                      | 1,042                             |                                   | food              | 28 days                | survival/ growth |                                       | Kamunde et al. (2001)     | non-significant growth inhibition at NOAEL; only no-effect level concentration reported; fish fed prepared diet   |
| Copper         | copper sulfate pentahydrate  | grey mullet       |                      |                                   | 2,397 <sup>b</sup>                | food              | 67 days                | growth           |                                       | Baker et al. (1998)       | effects associated with reduced feeding   |
| Copper         | copper sulfate               | rainbow trout     |                      | 10,000                            |                                   | food              | 28 days                | survival         |                                       | Handy (1993)              | only no-effect level concentration reported; fish fed prepared diet treated with chemical salts   |
| Lead           | lead nitrate                 | rainbow trout     | fry                  | 210 <sup>b</sup>                  |                                   | primarily food    | 60 days                | survival, growth | 100% survival, body weight and length | Mount et al. (1994)       | fish exposed to copper, cadmium, lead, and zinc in water at 23.0, 0.97, 3.32, and 46.3 µg/L in addition to dietary exposure; fish fed live <i>Artemia</i> exposed to lead nitrate in water  |
| Lead           |                              | rainbow trout     |                      | 7,040                             |                                   | food              |                        | growth           |                                       | Goettl et al. (1976)      | fish fed prepared diet treated with chemical salts  |
| Silver         |                              | rainbow trout     |                      | 3000                              |                                   | food              |                        | growth           |                                       | Galvez and Wood (1999)    | fish fed prepared diet treated with chemical salts  |
| Vanadium       | sodium orthovanadate         | rainbow trout     | juvenile             | 2.04 <sup>d</sup>                 | 10.2                              | food              | 12 wks                 | growth           | body weight                           | Hilton and Bettger (1988) | reduced feeding at all doses but significant increase in food consumption/weight gain   |
| Zinc           | zinc chloride                | rainbow trout     | fry                  | 1,900 <sup>b</sup>                |                                   | primarily food    | 60 days                | growth, survival | body weight and length; 97% survival  | Mount et al. (1994)       | fish exposed to copper, cadmium, lead, and zinc in water at 23.0, 0.97, 3.32, and 46.3 µg/L in addition to dietary exposure. fish fed live <i>Artemia</i> exposed to zinc chloride in water   |
| Zinc           |                              | rainbow trout     | fingerling           | 1,000                             | 2,000                             | food              | 6 wks                  | growth           |                                       | Takeda and Shimma (1977)  | fish fed at same dose zinc with 0.5% Ca experienced no adverse effects.   |
| Benzo(a)pyrene |                              | areolated grouper | juvenile             | 81                                |                                   | food              | 4 wks + 4 wks recovery | survival, growth | survival, body weight & length        | Wu et al. (2003)          | exposure was dietary force feeding - fish were force fed pellet into larynx   |
| Benzo(a)pyrene |                              | English sole      | juvenile             | 47                                | 116                               | food              | 10-12 days             | growth           |                                       | Rice et al. (2000)        | diet of exposed polychaetes exposed for 28 d; exposure concentrations calculated from feeding rates   |
| Benzo(a)pyrene |                              | rainbow trout     | juvenile             | 100                               | 1,000                             | food              | 28 days                | growth           |                                       | Hart and Heddle (1991)    | fish fed prepared diet treated with benzo(a)pyrene  |
| Benzo(a)pyrene |                              | rainbow trout     | 3 mo. old            |                                   | 1,000                             | food              | 18 mo                  | growth           | reduced body weight                   | Hendricks et al. (1985)   | fish fed prepared diet treated with benzo(a)pyrene; total mortality was higher in the control group   |
| Benzo(a)pyrene |                              | rainbow trout     | juvenile (3 mos old) | 1,000                             |                                   | food              | 18 mo                  | survival         |                                       | Hendricks et al. (1985)   | mortality was greater in control; NOAEL supported by additional 9-week study; fish fed prepared diet treated with benzo(a)pyrene  |

Table 1, continued

| CHEMICAL | CHEMICAL FORM            | TEST SPECIES   | LIFE STAGE | NOAEL (mg/kg dw <sup>a</sup> ) | LOAEL (mg/kg dw <sup>a</sup> ) | EXPOSURE ROUTE | EXPOSURE DURATION                      | ENDPOINT           | ENDPOINT EFFECT                | SOURCE               | NOTES   |
|----------|--------------------------|----------------|------------|--------------------------------|--------------------------------|----------------|--|--------------------|--------------------------------|----------------------|---|
| PAHs     | PAH mixture <sup>a</sup> | chinook salmon | juvenile   | 280 <sup>b</sup>               |                                | food           | 7 wks                                  | growth             |                                | Palm et al. (2003)   | 14 PAHs included in diet: acenaphthene, benzo(a)anthracene, benzo(b)fluoranthene, benzo(g,h,i)perylene, benzo(k)fluoranthene, chrysene, dibenzo(a,h)anthracene, fluoranthene, fluorene, indeno(1,2,3-cd)pyrene, naphthalene, phenanthrene, and pyrene   |
| PAHs     | PAH mixture <sup>a</sup> | chinook salmon | juvenile   | 280 <sup>b</sup>               |                                | food           | 4 wks exposure, 2 wks immuno-challenge | disease resistance |                                | Palm et al. (2003)   | 14 PAHs included in diet: acenaphthene, benzo(a)anthracene, benzo(b)fluoranthene, benzo(g,h,i)perylene, benzo(k)fluoranthene, chrysene, dibenzo(a,h)anthracene, fluoranthene, fluorene, indeno(1,2,3-cd)pyrene, naphthalene, phenanthrene, and pyrene; fish were exposed to <i>Listonella anguillarum</i> following PAH exposure. No difference was observed between PAH-exposed fish and controls in either fish that were vaccinated against the bacterium or those that weren't vaccinated |
| PAHs     | PAH mixture              | chinook salmon | juvenile   | 324                            | 951                            | food           | 53 days                                | growth             | reduced dry-weight body weight | Meador et al. (2006) | 21 PAHs included in diet: naphthalene, 2-methylnaphthalene, dimethylnaphthalene, dibenzothiophene, acenaphthene, fluorene, 1,8-dimethyl(9H)fluorene, phenanthrene, 9-ethylphenanthrene, 9-ethyl-10-methylphenanthrene, 1-methyl-7-isopropylphenanthrene, anthracene, fluoranthene, pyrene, methyl pyrene, benzo(a)anthracene, chrysene, benz(a)pyrene, benzo(k)fluoranthene, benzo(g,h,i)perylene, and dibenzanthracene   |

<sup>a</sup> Concentrations of elemental metal or specific PAH chemical. Except where noted, concentrations were as administered, and were assumed to approximate a dry weight concentration.

<sup>b</sup> Converted to dry weight basis based on moisture content reported in paper.

<sup>c</sup> NOAEL estimated using uncertainty factor of 5 (chronic LOAEL to chronic NOAEL).

<sup>d</sup> NOAEL estimated using uncertainty factor of 10 (acute LOAEL to chronic NOAEL).

LOAEL – lowest-observed-adverse-effect level

NOAEL – no-observed-adverse-effect level

PAH – polycyclic aromatic hydrocarbon

Table 2. Acceptable toxicity studies for selection of fish whole-body tissue residue TRVs

| CHEMICAL     | NOAEL (WB)       | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES   | LIFESTAGE                   | EXPOSURE MODE     | EXPOSURE DURATION        | ENDPOINT                    | ENDPOINT EFFECT                       | SOURCE                                     | NOTES  |
|--------------|------------------|------------|--------------|-------------|-------------|-----------|----------------|-----------------------------|-------------------|--------------------------|-----------------------------|---------------------------------------|--|--|
| 2,3,7,8-TCDD | 2.4 <sup>a</sup> | 12         | 28           |             |             | pg/g      | rainbow trout  | adult                       | food              | approx 300 days          | survival                    | reduced survival                      | Giesy et al. (2002)                        | LOAEL of 0.44 in fillet tissue measured at day 200 converted to concentration at approximately day 300 when mortality was observed based on recommendation from the author (Giesy 2006)  |
| 2,3,7,8-TCDD | 46               | 85         |              |             |             | pg/g      | lake whitefish | juvenile                    | food              | 30 days                  | growth                      | body weight                           | Fisk et al. (1997)                         | exposed for 30 days and effect measured for 180 days post-exposure; unclear when residues measured; assumed WB residues, however liver tissue possibly excluded; 29% mortality in control  |
| 2,3,7,8-TCDD | 87               | 102        | 2.56         | 34          | 40          | pg/g      | lake trout     | egg (converted to adult WB) | water             |                          | reproduction (egg exposure) | sac-fry survival                      | Walker et al. (1994)                       | study measured through routes of exposure for the egg (maternal transfer, water, injection)—egg residues not statistically different based on exposure route; significant mortality observed at LOAEL; adult concentration was estimated using egg:adult conversion factor of 2.56 based on Tietge et. al. (1998)                      |
| 2,3,7,8-TCDD | 38               | 102        | 2.56         | 15          | 40          | pg/g      | lake trout     | egg (converted to adult WB) | water             | 48 hrs                   | reproduction (egg exposure) | early life-stage survival             | Spitsbergen et al. (1991)                  | 22.5% cumulative mortality (egg, hatching, sac-fry) at LOAEL statistically significant; mortality at NOAEL (15.8%) not statistically different from control (18%); egg:adult conversion factor of 2.56 based on Tietge et. al. (1998)  |
| 2,3,7,8-TCDD | 90               | 108        | 2.56         | 35          | 42          | pg/g      | lake trout     | egg (converted to adult WB) | water             | 48 hrs                   | reproduction (egg exposure) | early life-stage survival             | Guiney et al. (1996)                       | NOAEL and LOAEL (LC50) based on hatchery fish eggs; LOAEL is LC50; NOAEL and LC50 also reported in field-collected eggs (from Lake Ontario) and reference eggs (from Lake Ontario); Lake Ontario NOAEL=30-45 pg/g and LC50=44-72 pg/g; Lake Superior NOAEL=34 pg/g; egg:adult conversion factor of 2.56 based on Tietge et. al. (1998) |
| 2,3,7,8-TCDD | 59               | 128        | 2.56         | 23          | 50          | pg/g      | lake trout     | egg (converted to adult WB) | maternal transfer | 11 wks prior to spawning | reproduction (egg exposure) | sac-fry survival                      | Walker et al. (1994)                       | parental fish exposed via diet; study measured through routes of exposure for the egg (maternal transfer, water, injection)—egg residues not statistically different based on exposure route; significant mortality observed at LOAEL; mortality in control=7.2%; egg:adult conversion factor of 2.56 based on Tietge et. al. (1998)   |
| 2,3,7,8-TCDD | 87               | 141        | 2.56         | 34          | 55          | pg/g      | lake trout     | egg (converted to adult WB) | water             | 48 hrs                   | reproduction (egg exposure) | early life-stage and sac-fry survival | Walker et al. (1991); Walker et al. (1992) | statistically significant effects observed at LOAEL, frequency of sac-fry edema also statistically significant; LC50=65 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |
| 2,3,7,8-TCDD | 72               | 150        |              |             |             | pg/g      | rainbow trout  | juvenile                    | food              | 30 days                  | growth                      | body weight                           | Fisk et al. (1997)                         | exposed for 30 days and effect measured for 180 days post-exposure; unclear when residues measured; assumed WB residues, however liver tissue possibly excluded  |
| 2,3,7,8-TCDD | 215              | 225        | 2.56         | 84          | 88          | pg/g      | brook trout    | egg (converted to adult WB) | maternal transfer | throughout egg dev'l'p   | reproduction (egg exposure) | swim-up and juvenile survival         | Johnson et al. (1998)                      | parental fish (1.5 yr old) fed TCDD until spawning; LOAEL is calculated LC10; LC10 is less than residue at which mortality was observed (at 156 pg/g ww); NOAEL is egg residue where no significant mortality was observed; egg residues are approximate; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)            |
| 2,3,7,8-TCDD | 346              | 474        | 2.56         | 135         | 185         | pg/g      | brook trout    | egg (converted to adult WB) | water             | 48 hrs                   | reproduction (egg exposure) | sac-fry survival                      | Walker and Peterson (1994)                 | NOAEL and LOAEL as reported in study; at LOAEL 25% mortality; at NOAEL 10% mortality; egg:adult conversion factor of 2.56 based on Tietge et al. (1988);   |
| 2,3,7,8-TCDD | 310              | 579        | 2.56         | 121         | 226         | pg/g      | lake trout     | egg (converted to adult WB) | water             | 48 hrs                   | reproduction (egg exposure) | sac-fry growth (length), hatchability | Walker et al. (1991)                       | statistically significant effects observed at LOAEL; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)   |

Table 2, continued

| CHEMICAL     | NOAEL (WB) | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES           | LIFESTAGE                   | EXPOSURE MODE | EXPOSURE DURATION                   | ENDPOINT  | ENDPOINT EFFECT                      | SOURCE                                     | NOTES   |
|--------------|------------|------------|--------------|-------------|-------------|-----------|------------------------|-----------------------------|---------------|-------------------------------------|---|--------------------------------------|--|---|
| 2,3,7,8-TCDD | 448        | 691        | 2.56         | 175         | 270         | pg/g      | lake herring           | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; six other fish species tested- lake herring was the most sensitive fish exposed to 2,3,7,8-TCDD; LC50=902 pg/g; LC10=509 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)   |
| 2,3,7,8-TCDD |            | 714        | 2.56         |             | 279         | pg/g      | rainbow trout          | egg (converted to adult WB) | water         | 48 hrs                              | reproduction (egg exposure)                           | sac-fry survival                     | Walker et al. (1992)                       | LC50s also reported in study; sac-fry mortality most sensitive endpoint; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |
| 2,3,7,8-TCDD |            | 780        |              |             |             | pg/g      | rainbow trout          | 10-15cm                     | water         | 2 and 6 hrs                         | growth  | body weight                          | Branson et al. (1985)                      | no statistics; LOAEL presented is based on tissue residues measured 28d after end of exposure, LOAEL is misleading; higher LOAEL should be used (max WB for dose, WB at end of exp, etc); tissue burdens assoc with same dose but measured earlier in exp were higher   |
| 2,3,7,8-TCDD |            | 980        |              |             |             | pg/g      | rainbow trout          | fry                         | water         | ~28 days                            | survival  |                                      | Mehrle et al. (1988)                       | LOAEL is based on tissue residue measured at the end of the exposure period   |
| 2,3,7,8-TCDD |            | 990        |              |             |             | pg/g      | rainbow trout          | fry                         | water         | ~28 days                            | growth  |                                      | Mehrle et al. (1988)                       |   |
| 2,3,7,8-TCDD | 602        | 1,114      | 2.56         | 235         | 435         | pg/g      | fathead minnow         | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; LC50=539 pg/g; LC10=293 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |
| 2,3,7,8-TCDD | 1,486      |            |              |             |             | pg/g      | brook trout            | adult                       | food          | 182 days                            | survival, growth, gonadal development, egg production |                                      | Tietge et al. (1998)                       | NOAEL is initial concentration in female fish exposed to 1,200 pg TCDD/g-food; Giesy et al. (2002) report adverse effects in rainbow trout at similar exposure levels after 250 days of exposure  |
| 2,3,7,8-TCDD |            | 1,626      | 2.56         |             | 635         | pg/g      | killifish (mummi-chog) | egg (converted to adult WB) | water         | through hatch or 30 days post-hatch | reproduction (egg exposure)                           | hatchability                         | Prince and Cooper (1995)                   | hatchability (61%); study compared the effects of TCDD-treated water on field-collected "non-impacted" eggs (from clean site - Tuckerton, NJ) and "impacted" eggs (from contaminated site - Newark Bay, NJ); no dose response from "impacted" eggs; egg:adult conversion factor of 2.56 based on Tietge et al. (1988) |
| 2,3,7,8-TCDD | 125        | 2,170      |              |             |             | pg/g      | coho salmon            | juvenile                    | water         | 96 hrs                              | growth  |                                      | Miller et al. (1979)                       | body burdens measured after 114 days in clean water following exposure, thus not conservative   |
| 2,3,7,8-TCDD | 125        | 2,170      |              |             |             | pg/g      | coho salmon            | juvenile                    | water         | 96 hrs                              | survival  |                                      | Miller et al. (1979)                       | body burdens measured after 114 days in clean water following exposure  |
| 2,3,7,8-TCDD | 986        | 2,189      | 2.56         | 385         | 855         | pg/g      | channel catfish        | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; LC50=644 pg/g; LC10=429 pg/g   |
| 2,3,7,8-TCDD |            | 2,410      |              |             |             | pg/g      | medaka                 | juvenile                    | water         | 12 days                             | growth  | body weight                          | Schmieder et al. (1995)                    | no statistics; growth increased 74% vs. 94% in control  |
| 2,3,7,8-TCDD | 1,165      | 2,429      | 2.56         | 455         | 949         | pg/g      | medaka                 | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; LC50=1,110 pg/g; LC10=656 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |
| 2,3,7,8-TCDD | 2,171      | 3,123      | 2.56         | 848         | 1,220       | pg/g      | white sucker           | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; LC50=1,890 pg/g; LC10=1,590 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |
| 2,3,7,8-TCDD | 3,046      | 4,608      | 2.56         | 1,190       | 1,800       | pg/g      | northern pike          | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; LC50=2,460 pg/g; LC10=1,530 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |
| 2,3,7,8-TCDD | 1,085      | 5,120      | 2.56         | 424         | 2,000       | pg/g      | zebra fish             | egg (converted to adult WB) | water         | 100 days                            | reproduction (egg exposure)                           | early life-stage growth and survival | Elonen et al. (1998); Spehar et al. (1997) | statistically significant effects observed at LOAEL; LC50=2,610 pg/g; LC10=1,610 pg/g; egg:adult conversion factor of 2.56 based on Tietge et al. (1988)  |

Table 2, continued

| CHEMICAL                              | NOAEL (WB) | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES      | LIFESTAGE             | EXPOSURE MODE     | EXPOSURE DURATION                               | ENDPOINT                    | ENDPOINT EFFECT  | SOURCE                   | NOTES   |
|---------------------------------------|------------|------------|--------------|-------------|-------------|-----------|-------------------|-----------------------|-------------------|---|-----------------------------|--|--------------------------|---|
| 2,3,7,8-TCDD                          |            | 69,000     |              |             |             | pg/g      | fathead minnow    | juvenile              | water             | 28 day  | survival                    |  | Adams et al. (1986)      | no statistics—100% mort; LOAEL is average concentration in dead fish  |
| 2,3,7,8-TCDD                          | 1,570      | 1,380,000  |              |             |             | pg/g      | rainbow trout     | young                 | food              | 105 days  | survival                    |  | Hawkes and Norris (1977) |   |
| 2,3,7,8-TCDD                          | 1,570      | 1,380,000  |              |             |             | pg/g      | rainbow trout     | young                 | food              | 105 days  | growth                      | body weight  | Hawkes and Norris (1977) |   |
| PCBs (Aroclor 1260)                   |            | 520        |              |             |             | µg/kg     | common barbel     | adult                 | food              | 50 days   | reproduction                | fecundity  | Hugla and Thome (1999)   | Fecundity effect was not dose-responsive; number of fish tested unclear; uncertain statistical significance of fecundity endpoint; ww concentration converted from dw assuming 20% solids   |
| PCBs (Aroclor mixture)                |            |            |              |             | 857         | µg/kg     | Atlantic salmon   | embryo                | water             | 48 hrs  | reproduction (egg exposure) | live fry body weight   | Fisher et al. (1994)     | no effect on reproduction were observed   |
| PCBs (Aroclor 1254)                   | 980        |            |              |             |             | µg/kg     | chinook salmon    | juvenile              | food              | 4 wks   | growth, survival            |  | Powell et al. (2003)     | whole body burdens ranged from 740 to 980 µg/kg ww over the 13-day period following treatment; only no-effect level reported; no effect on growth, survival, or survival following immunological challenge  |
| PCBs (Aroclor 1254; egg) <sup>a</sup> |            |            |              |             | 1,640       | µg/kg     | rainbow trout     | egg                   | maternal transfer | 60 days   | reproduction (egg exposure) | fry growth   | Hendricks et al. (1981)  | eggs were exposed via maternal transfer from single gravid female fed 200 µg/g PCBs for 60 days   |
| PCBs (Aroclor 1260)                   |            | 2,640      |              |             |             | µg/kg     | common barbel     | adult                 | food              | 75 days   | reproduction                | lack of spawning in first reproductive season; egg and larval mortality; fecundity | Hugla and Thome (1999)   | fecundity effect was not dose-responsive; number of fish tested unclear; uncertain statistical significance of fecundity endpoint; ww concentration converted from dw assuming 20% solids   |
| PCBs (Aroclor 1254)                   |            |            |              |             | 3,200       | µg/kg     | Atlantic croaker  | egg                   | maternal transfer | 2 wks during reproduction (adults)              | reproduction (egg exposure) | reduction in larval growth rate  | McCarthy et al. (2003)   | parental fish fed dietary PCBs, eggs exposed via maternal transfer; residues not clearly presented  |
| PCBs (Aroclor 1254)                   | 8,000      |            |              |             |             | µg/kg     | rainbow trout     | 14 wks                | food              | 32 wks  | growth, survival            |  | Lieb et al. (1974)       | only no-effect level reported   |
| PCBs (Aroclor 1254)                   | 1,900      | 9,300      |              |             |             | µg/kg     | sheepshead minnow | adult                 |                   | 28 days   | reproduction                | decreased fry survival   | Hansen et al. (1974a)    | elevated PCB concentrations in control fish; egg production was artificially stimulated by injecting the fish with human chorionic gonadotropic hormone   |
| PCBs (Aroclor 1268)                   | 15,000     |            |              |             |             | µg/kg     | mummichog         | adult                 | food              | ~6 wks  | reproduction                | fertilization and hatching success, larval survival                                | Matta et al. (2001)      | two generations of progeny observed; only no-effect level reported; offspring weight was significantly greater for PCB-exposed fish with parental tissue burdens equal to or greater than 1,300 µg/kg ww; however, this was not considered an adverse effect    |
| PCBs (Aroclor 1254)                   | 27,000     | 46,000     |              |             |             | µg/kg     | spot              |                       | water             | 20 days   | survival                    |  | Hansen et al. (1971)     | mortality did not appear directly related to body burden; bb increased with exposure duration; NOAEL (catfish) = 32   |
| PCBs (Aroclor 1254)                   | 31,000     | 71,000     |              |             |             | µg/kg     | brook trout       | fry- exposure to eggs | water             | 10 days prior to hatch and 118 days after hatch | growth                      | reduced growth   | Mauck et al. (1978)      | residue measured at 118 days but was absent at 118 days.  |
| PCBs (Aroclor 1254)                   |            |            |              |             | 77,900      | µg/kg     | brook trout       | egg                   | water             | 21 days   | reproduction (egg exposure) | reduced hatchability   | Freeman and Idler (1975) | 75% hatching at LOAEL and 92% hatching in control; concentration in back muscle of dose fish with affected hatchability was 32.8 mg/kg ww   |
| PCBs (Aroclor 1016)                   | 110,000    |            |              |             |             | µg/kg     | sheepshead minnow | fry                   | water             | 4 wks   | reproduction                | fertilization and hatching success, larval survival                                | Hansen et al. (1975)     | intermittent-flow toxicity test; no effect: fertilization success, survival of embryos to hatching, or survival of fry; only no-effect level reported; Juvenile tissue concentration LOAEL for reduced survival at the same exposure level was 220,000 µg/kg ww |

Table 2, continued

| CHEMICAL                           | NOAEL (WB) | LOAEL (WB)       | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES              | LIFESTAGE             | EXPOSURE MODE                    | EXPOSURE DURATION                             | ENDPOINT                    | ENDPOINT EFFECT                                   | SOURCE                    | NOTES  |
|------------------------------------|------------|------------------|--------------|-------------|-------------|-----------|---------------------------|-----------------------|----------------------------------|---|-----------------------------|---|---------------------------|--|
| PCBs (Aroclor 1016)                |            | 106,000          |              |             |             | µg/kg     | pinfish                   |                       | water                            | 33 days                                       | survival, behavior          | loss of equilibrium; erratic swimming             | Hansen et al. (1974b)     | significant reduction in survival (50% mortality relative to 6% in control)  |
| PCBs (Aroclor 1254:1260 mixture)   | 120,000    |                  |              |             |             | µg/kg     | rainbow trout             | young                 | water                            | 90 days                                       | survival                    |   | Mayer et al. (1985)       | mortality observed; not significantly different; dose was 1:2 ratio of Aroclor 1254:1260; only no-effect level reported  |
| PCBs (Aroclor 1254:1260 mixture)   | 70,000     | 120,000          |              |             |             | µg/kg     | rainbow trout             | young                 | water                            | 90 days                                       | growth                      |   | Mayer et al. (1985)       |  |
| PCBs (Aroclor 1254)                | 71,000     | 125,000          |              |             |             | µg/kg     | brook trout               | fry- exposure to eggs | water                            | 10days prior to hatch and 118days after hatch | survival                    | fry survival                                      | Mauck et al. (1978)       | reduced fry survival; 21 to 100% mortality; tissue residue measured at 118 days; Median hatching time and egg hatchability were not affected; larval growth was initially reduced, but not by the end of the test  |
| PCBs (Aroclor 1016)                | 57,000     | 200,000          |              |             |             | µg/kg     | sheepshead minnow         | fry                   | water                            |   | survival                    | fry survival                                      | Hansen et al. (1975)      |  |
| PCBs (Clophen A50)                 |            | 250,000          |              |             |             | µg/kg     | goldfish                  |                       | water                            | 5-21 days                                     | survival                    | survival  | Hattula and Karlog (1972) | LOAEL is concentration in individual dead fish   |
| PCBs (Aroclor 1254)                |            | 429,000 (female) |              |             |             | µg/kg     | fathead minnow            |                       | water                            |   | reproduction                | reduced spawning                                  | Nebeker et al. (1974)     | mean terminal residue; egg hatchability and fry survival was not affected  |
| PCBs (Aroclor 1242, 1254, or 1260) |            | 1,860 – 749,000  |              |             |             | µg/kg     | fathead minnow            |                       | water                            | up to 300 hrs                                 | survival                    | lethal body burden                                | van Wezel et al. (1995)   | Tissue concentrations of individual fish that died in less than 20 hours ranged from 1,800 to 30,000 µg/kg ww; tissue concentrations of individual fish that died at 100 to 300 hours ranged from 120,000 to 749,000 µg/kg ww.   |
| Mercury (mercuric chloride)        | 0.2        |                  |              |             |             | mg/kg     | guppy                     | male adult            | sediment and water               | 20 days                                       | survival                    |   | Kudo and Mortimer (1979)  | only no-effect level reported  |
| Mercury                            |            |                  |              |             | 14          | µg/kg     | catfish                   | 4 days post-hatch     | water                            | spawning until 4 d post-hatch                 | survival                    | 67% survival                                      | Birge et al. (1979)       | LOAEL is sac-fry concentration at 4 days post-hatch; control data was not reported; 17% survival was reported in catfish with WB residues of 1.15 µg/g ww  |
| Mercury (inorganic Hg)             |            |                  |              |             | 36          | µg/kg     | rainbow trout             | alevins               | exposed to contaminated sediment | 20 days                                       | reproduction (egg exposure) | egg/embryo survival reduced by 46%                | Birge et al. (1979)       | LOAEL is based on alevin (sac-fry) concentration of 0.036; adult concentration was estimated using sac-fry:adult conversion factor of 64.4 based on rainbow trout data from Niimi (1983) and assuming 3:2 ratio of concentration of sac-fry:egg  |
| Mercury                            |            |                  |              |             | 41          | µg/kg     | rainbow trout             | 4 days post-hatch     | water                            | spawning until 4 days post-hatch              | survival                    | 45% survival                                      | Birge et al. (1979)       | LOAEL is sac-fry concentration at 4 days post-hatch; 94% survival was reported in controls; 39 - 79% survival was reported in eggs with concentrations of 0.0682 µg/g ww; adult concentration was estimated using sac-fry:adult conversion factor of 64.4 based on rainbow trout data from Niimi (1983) and assuming 3:2 ratio of concentration of sac-fry:egg |
| Mercury (methylmercury)            | 230        |                  |              |             |             | µg/kg     | golden shiner             |                       | food                             | 90 days                                       | survival                    | predator avoidance                                | Webber and Haines (2003)  | Field collected fish (but from a manmade cement lined pond). Predator avoidance was reduced at this NOAEL; growth accompanied by reduced feeding was reduced at NOAEL but not dose-responsive, no effect on growth was observed in fish fed 959 ng/g ww where WB residue = 518 ug/g ww   |
| Mercury (methylmercuric chloride)  | 200        | 470              |              |             |             | µg/kg     | mummichog                 | adult                 | water                            | 42 days                                       | survival                    |   | Matta et al. (2001)       | effects observed in male fish but not females  |
| Mercury (methylmercury)            |            |                  |              | 990         |             | µg/kg     | walleye (field-collected) | egg                   | maternal transfer                |   | reproduction (egg exposure) | fertilization and hatching success, larval length | Latif et al. (2001)       | field-collected fish; adult concentration was estimated using egg:adult conversion factor of 36.05 based on average data reported in five species in Niimi (1983); mercury concentrations of diet not reported   |
|                                    |            |                  |              |             |             |           |                           |                       |                                  |   |                             |   |                           |  |
| Mercury (mercuric chloride)        | 0.8        | 1,310            |              |             |             | µg/kg     | fathead minnow            | 3 mos                 | food                             | 60 days                                       | growth                      |   | Snarski and Olson (1982)  |  |



Table 2, continued

| CHEMICAL                           | NOAEL (WB) | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES   | LIFESTAGE                       | EXPOSURE MODE | EXPOSURE DURATION           | ENDPOINT         | ENDPOINT EFFECT   | SOURCE                      | NOTES   |
|------------------------------------|------------|------------|--------------|-------------|-------------|-----------|----------------|---------------------------------|---------------|-----------------------------|------------------|---|-----------------------------|---|
| Mercury (methyl-mercuric chloride) | 2,280      |            |              |             |             | µg/kg     | rainbow trout  | fingerling                      | water         | 24 days                     | growth           |   | Phillips and Buhler (1978)  | only no-effect level reported   |
|                                    |            |            |              |             |             |           |                |                                 |               |                             |                  |   |                             |   |
|                                    |            |            |              |             |             |           |                |                                 |               |                             |                  |   |                             |   |
| Mercury (methylmercury)            | 2,700      | 3,400      |              |             |             | µg/kg     | brook trout    | embryo – adult                  | water         | 756 days                    | reproduction     | reduced number of viable eggs produced                              | McKim et al. (1976)         | residue measured in parental fish at 39 weeks; NOAEL is recommended no effect level presented in study (for abnormalities and multigenerational mortalities); NOAEL based on the data for no effect on the number of viable eggs produced was 1.1 µg/g ww |
| Mercury (methylmercury)            |            |            |              |             | 3,800       | µg/kg     | grayling       | sac-fry (converted to adult WB) | water         | 1st 10 days of develop-ment | survival         | increased fry mortality   | Fjeld et al. (1998)         | NOAEL and LOAEL reported as yolk-fry concentration; adult concentration was estimated using sac-fry:adult conversion factor of 64.4 based on rainbow trout data from Niimi (1983) and assuming 3:2 ratio of concentration of sac-fry:egg                  |
| Mercury (mercuric chloride)        | 2,750      | 4,180      |              |             |             | µg/kg     | fathead minnow | 3 month old                     | water         | 60 days                     | survival         |   | Snarski and Olson (1982)    |   |
| Mercury (mercuric chloride)        | 2,840      | 4,470      |              |             |             | µg/kg     | fathead minnow | larvae-adult                    | water         | 287 days                    | reproduction     |   | Snarski and Olson (1982)    |   |
| Mercury (methylmercury)            | 5,000      |            |              |             |             | µg/kg     | rainbow trout  | juvenile                        | water         | 84 days                     | growth, survival |   | Lock (1975)                 | only no-effect level reported   |
| Mercury (mercuric chloride)        |            | 5,600      |              |             |             | µg/kg     | goldfish       | 4.5-6.5 cm                      | water         | 2 days                      | survival         |   | Heisinger et al. (1979)     | Concentration converted from dry weight to wet weight assuming 80% moisture in whole fish   |
| Mercury (methyl-mercuric chloride) | 5,670      |            |              |             |             | µg/kg     | rainbow trout  | fingerling                      | water         | 24 days                     | growth           |   | Phillips and Buhler (1978)  | only no-effect level reported   |
| Mercury (methyl-mercuric chloride) |            | 6,500      |              |             |             | µg/kg     | bluegill       | juvenile                        | water         | 12.5 days                   | survival         |   | Cember et al. 1978          |   |
| Mercury (methyl-mercuric chloride) | 8,630      |            |              |             |             | µg/kg     | rainbow trout  | fingerling                      | water         | 24 days                     | growth           |   | Phillips and Buhler (1978)  | brook trout = 9.4 (NOAEL); only no-effect level reported  |
| Mercury                            | 9,400      |            |              |             |             | µg/kg     | brook trout    | 39 wks old                      | water         | 756 days                    | growth, survival | reduced juvenile weight   | McKim et al. (1976)         | residues measured in parental fish at 39 weeks; only no-effect level reported   |
| Mercury (methylmercury)            |            | 10,000     |              |             |             | µg/kg     | rainbow trout  | fingerling                      | food          | 84 days                     | growth           | reduced growth and final body weight (after 84 days)                | Rodgers and Beamish (1982)  | reduced growth associated with feeding; approximated from graph (and text); at 14 days of treatment Hg concentrations in WB tissues were lower (approx. 5 µg/g ww); however, effects were observed (and significant) over the 84 days of exposure         |
| Mercury (methylmercury)            | 10,900     |            |              |             |             | µg/kg     | fathead minnow | larvae-adult                    | water         | 336 days                    | growth, survival |   | Olson et al. (1975)         | only no-effect level reported   |
| Mercury (methyl-mercuric chloride) | 1,100      | 11,000     |              |             |             | µg/kg     | mummichog      |                                 | food          | 42 days                     | reproduction     | F1 fertilization success  | Matta et al. (2001)         | residue measured in parental fish   |
| Mercury (methyl-mercuric chloride) | 12,000     |            |              |             |             | µg/kg     | rainbow trout  | sub-adult                       | water         | 75 days                     | growth, survival | body weight   | Niimi and Lowe-Jinde (1984) | only no-effect level reported   |
| Mercury (methyl-mercuric chloride) | 12,000     |            |              |             |             | µg/kg     | mummichog      |                                 | food          | 42 days                     | reproduction     | F1 hatchability, survival, fecundity, F2 larval survival            | Matta et al. (2001)         | residue measured in female parental fish; only no-effect level reported   |
| Mercury (methyl-mercuric chloride) | 12,000     |            |              |             |             | µg/kg     | mummichog      |                                 | food          | 42 days                     | reproduction     | fecundity, fertilization success, offspring weight, female survival | Matta et al. (2001)         | offspring of medium and high doses were larger than offspring of control fish; only no-effect level reported; residue measured in female parental fish  |
| Mercury                            | 29,000     |            |              |             |             | µg/kg     | rainbow trout  | fingerling                      |               | 84 days                     | survival         |   | Rodgers and Beamish (1982)  | only no-effect level reported   |

Table 2, continued

| CHEMICAL  | NOAEL (WB) | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES      | LIFESTAGE                   | EXPOSURE MODE  | EXPOSURE DURATION                         | ENDPOINT                    | ENDPOINT EFFECT  | SOURCE                     | NOTES  |
|---|------------|------------|--------------|-------------|-------------|-----------|-------------------|-----------------------------|----------------|---|-----------------------------|--|----------------------------|--|
| Mercury (methylmercury)                         |            |            |              | 16,000      | 29,000      | µg/kg     | Japanese medaka   | egg (converted to adult WB) | water          | 16 days                                   | reproduction (egg exposure) | hatchability   | Heisinger and Green (1975) | at LOAEL - 20.8% hatchability; at NOAEL - 58.3% hatchability; at control - 46.7% hatchability; adult concentration was estimated using egg:adult conversion factor of 36.05 based on average data reported in five species in Niimi (1983)                       |
| Selenium  | 1,200      | 1,600      |              |             |             | µg/kg     | National criteria | na                          | na             | na  | protection of aquatic life  | chronic criteria   | EPA (2004)                 | concentration converted to ww assuming 80% moisture content;   |
| TBT   | 18         | 159        |              |             |             | µg/kg     | Japanese flounder | larvae                      | food           | approx 65 days                            | growth                      | body weight  | Shimasaki et al. (2003)    | all test fish were genetically XX fish by parents both XX but one parent phenotypically male and functional; No replication in study; survival was not significantly affected at any dose; mortality was observed in all groups, including the control group.    |
| TBT   |            | 1,054      |              |             |             | µg/kg     | Japanese medaka   | adult                       | food           | 3 wks                                     | reproduction                | reduced swim-up failure  | Nakayama et al. (2005)     | LOAEL is concentration in adult female fish using an adult:egg conversion factor of 8.57 based on Nirmala et al. (1999); effects were observed in offspring; concentration in egg tissues was 123 µg/kg  |
| TBT   |            | 2,390      |              |             |             | µg/kg     | Japanese medaka   | adult                       | food           | 3 wks                                     | reproduction                | reduced hatching success, swim-up success, and embryonic success | Nirmala et al. (1999)      | LOAEL is concentration in adult female fish; effects were observed in offspring  |
| Chlordane (alpha-chlordane and gamma-chlordane) | 190 - 710  |            |              |             |             | µg/kg     | goldfish          |                             | food and water | 96 hrs                                    | survival                    | survival   | Moore et al. (1977)        | uncontrolled study; NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported   |
| Chlordane (cis-chlordane)                       |            | 1,360      |              |             |             | µg/kg     | goldfish          |                             | water          | 24 hrs                                    | survival                    | survival 3/9 fish died   | Feroz and Khan (1979)      | uncontrolled study; no clear residue burden and response to mortality effects; LOAEL is based on residues in carcasses of 3 dead fish; surviving fish tissue residues were 2.9 µg/g ww (at 10 days after exposure) and 2.1 µg/g (at 25 days after exposure)      |
| Chlordane (tech chlordane)                      |            | 16,600     |              |             |             | µg/kg     | pinfish           | fry/ juvenile               | water          | 96 hrs                                    | survival                    | 30% mortality  | Parrish et al. (1976)      | 96-hr exposure; tissue residues in surviving fish exposed to 5.4 µg/L aqueous chlordane; seawater  |
| Chlordane (tech chlordane)                      | 87,000     |            |              |             |             | µg/kg     | sheepshead minnow | fry/juvenile                | water          | 28 day                                    | survival                    | 3.7% mortality   | Parrish et al. (1976)      | 28-day embryo/fry exposure; tissue residues in surviving fish exposed to 7.1 µg/L; tissue residues not reported in fish exposed to higher aqueous concentrations (100% mortality); seawater  |
| Chlordane (tech chlordane)                      |            | 281,000    |              |             |             | µg/kg     | sheepshead minnow | fry/juvenile                | water          | 96 hrs                                    | survival                    | 25% mortality  | Parrish et al. (1976)      | 96-hr exposure; tissue residues in surviving fish exposed to 15 µg/L; seawater; tissue residue concentration at lowest water treated concentration (unbounded)   |
| DDT (mixture)                                   | 25         |            |              |             |             | µg/kg     | golden shiner     |                             | food           | 6-15 days                                 | survival                    | survival   | Courtney and Reed (1971)   | converted from dry weight to wet weight using factor given in paper  |
| DDT (mixture)                                   |            | 550        |              |             |             | µg/kg     | pinfish           | 4.8 g                       | diet           | 15 days                                   | survival                    | 44% mortality at 10 days   | Butler (1969)              | 4% mortality in control; residues were much higher in live fish (LOAEL is residue of dead fish); at 7 days live fish WB residue = 2,700 µg/kg ww; residues were 640 µg/kg ww where morality was 10% (at 14 days); poorly reported study; no clear dose-response  |
| DDT (mixture)                                   | 1,800      | 1,800      |              |             |             | µg/kg     | cutthroat trout   | 21 months                   | water          | 612 days (mortality observed at 111 days) | survival                    |  | Allison et al. (1964)      | LOAEL is tissue concentration at 111 days (3.7 months) in fish exposed to 0.1 mg/kg DDT in water where mortality was significant after "4 months" (approximately 120d), note that tissue concentrations at this dose increased to 3.0 mg/kg at the next sampling |
| DDT (total)                                     | 1,920      |            |              |             |             | µg/kg     | brook trout       | juvenile                    |                | 120 days                                  | survival                    |  | Macek and Korn (1970)      | no effect level reported   |

Table 2, continued

| CHEMICAL      | NOAEL (WB) | LOAEL (WB)      | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES                 | LIFESTAGE         | EXPOSURE MODE | EXPOSURE DURATION                  | ENDPOINT         | ENDPOINT EFFECT                             | SOURCE                       | NOTES  |
|---------------|------------|-----------------|--------------|-------------|-------------|-----------|------------------------------|-------------------|---------------|------------------------------------|------------------|---|------------------------------|--|
| DDT (total)   | 1,200      | 2,000           |              |             |             | µg/kg     | cutthroat trout              | yearlings         | water         | 20 mo (mortality observed at 4 mo) | survival         | survival                                    | Allison et al. (1963)        | LOAEL is tissue concentration at 56 days (1.9 months) in fish exposed to 1.0 mg/kg DDT in water where mortality was significant after "4 months" (approximately 120d), note that tissue concentrations at this dose increased to 4,300 µg/kg at the next sampling.   |
| DDT (total)   |            | 2,800 - 3,000   |              |             |             | µg/kg     | brook trout                  | sex mat yearlings | food          | 156 days                           | reproduction     | offspring (sac-fry and embryo) mortality    | Macek (1968b)                | LOAEL is range of parental tissue residues; residues in offspring (fry) ranged from 2,620 to 3,090 µg/kg ww  |
| DDT (total)   | 620        | 3,650           |              |             |             | µg/kg     | chinook salmon               | 0.61g             | food          | 40 days                            | survival         |   | Buhler et al. (1969)         | tissue concentration at 4 and 7 days; growth increase attributed to size -selective mortality  |
| DDT (total)   | 4,670      |                 |              |             |             | µg/kg     | rainbow trout                | 15 g              | food          | 140 days                           | growth, survival | survival, growth                            | Macek et al. (1970)          | no effect level reported   |
| DDT (total)   |            | 5,200           |              |             |             | µg/kg     | killifish                    |                   | water         | 24 hrs                             | survival         | 25% mortality at 24 hrs                     | Crawford and Guarino (1976)  | LOAEL is the weighted sum of residues in 10 different tissues of fish exposed to 0.1 ppm DDT for two 24hr DDT doses at 24 hrs after exposures; at 8 days after exposures, weighted sum of residues was 3,900 µg/kg ww, however, mortality started occurring at 24 hours.   |
| DDT (mixture) | 3,900      | 5,500           |              |             |             | µg/kg     | cutthroat trout              | 21 month old      | food          | 612 days                           | survival         | survival                                    | Allison et al. (1964)        | control diet had 760 µg/kg DDT; LOAEL tissue residue measured at day 166 in fish fed to 1,000 µg/kg body weight (bw) DDT and where mortality was observed; NOAEL is highest tissue concentration in fish fed 300 µg/kg bw DDT where significant mortality was not observed over entire exposure duration of 20 months; concentration is total chlorinated hydrocarbons |
| DDT (mixture) |            | 5,600           |              |             |             | µg/kg     | pinfish                      | 3.0 g             | food          | 21 days                            | survival         | 35% mortality at 21 days                    | Butler (1969)                | 0% mortality in control; LOAEL is residue of dead fish; residues were 3,300 µg/kg ww where morality was 10% (at 14 days); poorly reported study; no clear dose-response of tissues with mortality across the experiments presented in the paper  |
| DDT (mixture) |            | 7,890           |              |             |             | µg/kg     | pinfish                      | 3.0 g             | food          | 2 days                             | survival         | 63.5% mortality                             | Butler (1969)                | 0% mortality in control; LOAEL is residue of dead fish; poorly reported study; no clear dose-response of tissues with mortality across the experiments presented in the paper  |
| DDT (total)   | 7,600      |                 |              |             |             | µg/kg     | brook trout fry              | fry               | food          | 156 days                           | growth, survival | survival, growth                            | Macek (1968b)                | other fish species from study not reported because NOAEL/LOAELs are much higher; no effect level reported; length but not weight of male fish was significantly greater than controls  |
| DDT (total)   | 11,200     |                 |              |             |             | µg/kg     | brook trout                  | under yearling    | food          | 31 wks                             | growth           | increased growth                            | Macek (1968a)                | NOAEL effect is an increase in growth (significant); sum of DDD, DDE, and DDT concentrations   |
| DDT (total)   | 11,400     | 12,100          |              |             |             | µg/kg     | chinook salmon               | 1.1 g             | food          | 40 days                            | survival         | survival                                    | Buhler et al. (1969)         | tissues sampled at 4 and 7 days  |
| DDT (total)   |            | 20,200 - 45,800 |              |             |             | µg/kg     | brook trout                  | under yearling    | food          | 26 wks                             | survival         | reduced survival during stress (starvation) | Macek (1968a)                | sum of DDD, DDE, 2,4'-DDT and 4,4'-DDT in fish exposed to 2.0 mg/kg/wk DDT   |
| DDT (total)   |            | 24,000          |              |             |             | µg/kg     | green sunfish/ pumpkin--seed |                   | water         | 90 days                            | survival         | survival                                    | Hamelink et al. (1971)       |  |
| DDT (mixture) | 24,000     |                 |              |             |             | µg/kg     | Atlantic menhaden            |                   | food          | 48 (109) <sup>a</sup>              | growth           | growth                                      | Warlen et al. (1977)         | fish exposed for 48 days and observed for 109 days after exposure; no effect level reported  |
| DDT (total)   |            | 26,500          |              |             |             | µg/kg     | mosquito fish                |                   | water         | 16 days                            | survival         | survival                                    | Pillai et al. (1977)         |  |
| DDT (total)   | 40,000     |                 |              |             |             | µg/kg     | fathead minnow               | juvenile-adult    |               | 266 days                           | survival         | survival                                    | Jarvinen et al. (1976; 1977) | no effect level reported   |

Table 2, continued

| CHEMICAL  | NOAEL (WB)        | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES      | LIFESTAGE   | EXPOSURE MODE  | EXPOSURE DURATION             | ENDPOINT         | ENDPOINT EFFECT                                 | SOURCE                   | NOTES  |
|---|-------------------|------------|--------------|-------------|-------------|-----------|-------------------|-------------|----------------|-------------------------------|------------------|---|--------------------------|--|
| DDT (total)   | 51,400            | 92,700     |              |             |             | µg/kg     | sailfin molly     |             | water          | 21 days                       | growth, survival | survival, growth                                | Benton et al. (1994)     | converted from dry weight to wet weight based on reported 24% solids; sum of 2,4'-DDT, 4,4'-DDT, 2,4'-DDE, 4,4'-DDE, 2,4'-DDD, and 4,4'-DDD  |
| DDT (total)   | 16,600            | 69,600     |              |             |             | µg/kg     | coho salmon       | 3.7 g       | food           | 60 days                       | survival         | survival  | Buhler et al. (1969)     | tissues sampled at 4 and 7 days  |
| DDT (total)   | 130,000           |            |              |             |             | µg/kg     | goldfish          | 9.1 g       | food and water | 58 days                       | survival         | survival  | Rhead and Perkins (1984) | converted from dry weight to wet weight assuming 20% solids; no effect level reported  |
| DDT (total)   |                   | 200,000    |              |             |             | µg/kg     | goldfish          | 9.1 g       | food and water | 38 days                       | survival         | survival  | Rhead and Perkins (1984) | converted from dry weight to wet weight assuming 20% solids  |
| Dieldrin  | 85 - 120          |            |              |             |             | µg/kg     | largemouth bass   | adult       | food           | 30-50 days                    | survival, growth | body weight, length, condition factor, survival | Muller et al. (2004)     | estimated from graph; not clear if gonads included (measured at 1-3.5% of whole body concentration); blood not included  |
| Dieldrin  | 120               | 200        |              |             |             | µg/kg     | rainbow trout     | juvenile    | food and water | 16 wks                        | survival         | 33% mortality                                   | Shubat and Curtis (1986) | 5 of 15 fish died in high treated water (no maintenance diet treatment) and 4 of 15 fish died in high treated water and maintenance treated diet; no statistics; assumed no mortalities in control; no replication |
| Dieldrin  | 1,400             |            |              |             |             | µg/kg     | rainbow trout     | juvenile    |                | 16 wks                        | growth           |   | Shubat and Curtis (1986) | highest residue concentration in growth diet group (maintenance diet group with residues up to 0.36 ppm); no effect observed in body weight of treated fish groups   |
| Endosulfan (tech endosulfan (sum of endosulfan I, II, and endosulfan sulfate) | 3.1 <sup>b</sup>  | 31         |              |             |             | µg/kg     | spot              |             | water          | 96 hrs                        | survival         | 35% mortality; 10% mortality (control)          | Schimmel et al. (1977a)  | tissue residues of surviving fish  |
| Endosulfan (tech endosulfan (sum of endosulfan I, II, and endosulfan sulfate) | 195               | 272        |              |             |             | µg/kg     | pinfish           |             | water          | 96 hrs                        | survival         | 35% mortality LOAEL; 5% NOAEL                   | Schimmel et al. (1977a)  | tissue residues of surviving fish  |
| Endosulfan (tech endosulfan (sum of endosulfan I, II, and endosulfan sulfate) |                   | 360        |              |             |             | µg/kg     | mullet            |             | water          | 96 hrs                        | survival         | 40% mort; 0% mortality (control)                | Schimmel et al. (1977a)  | tissue residues of surviving fish  |
| Endrin (technical endrin)   | 1.15 <sup>b</sup> | 11.5       |              |             |             | µg/kg     | largemouth bass   | fingerlings | water          | 20 days                       | survival         | 40% fingerling mortality                        | Fabacher (1976)          | one dose; no statistics; static tanks treated every 5 days; residues in dead fish  |
| Endrin (analytical endrin)  |                   | 240        |              |             |             | µg/kg     | fathead minnow    | 30d         | food and water | 300 days (also reprod period) | survival         |   | Jarvinen and Tyo (1978)  | statistically significant results at LOAEL   |
| Endrin  | 307               |            |              |             |             | µg/kg     | channel catfish   | fingerlings | food           | 198 days                      | survival         |   | Argyle et al. (1973)     | no statistics; only no-effect level reported   |
| Endrin  | 307               |            |              |             |             | µg/kg     | channel catfish   | fingerlings | food           | 198 days                      | growth           |   | Argyle et al. (1973)     | no statistics; only no-effect level reported; LOAEL is average residue from day 20 to day 198  |
| Endrin  | 410               | 720        |              |             |             | µg/kg     | channel catfish   | fingerlings | water          | 55 days                       | survival         |   | Argyle et al. (1973)     | no statistics - 40% mortality; LOAEL is concentration at day 26 when mortality began; NOAEL is mean concentration when residue levels were maximal from days 49-55   |
| Endrin  | 110               | 880        |              |             |             | µg/kg     | sheepshead minnow | all         | water          | 4 wks                         | survival         | juvenile survival- 1 month survival of F2       | Hansen et al. (1977)     | saltwater species, embryos spawned from field-collected adults; tissue residues in juvenile fish- effects based on mortality effects in fry  |
| Endrin  | 260               | 940        |              |             |             | µg/kg     | sheepshead minnow | all         | water          | 2 generations                 | reproduction     | female fertility                                | Hansen et al. (1977)     | saltwater species, embryos spawned from field-collected adults   |
| Endrin  |                   | 1,660      |              |             |             | µg/kg     | golden shiner     |             | water          | 8 hrs                         | survival         | 100% mortality                                  | Ludke et al. (1968)      | LOAEL is average residue at time of death; average residue over 8 hours = 0.97 mg/kg ww (resistant [i.e. pre-exposed] test fish not evaluated)   |

Table 2, continued

| CHEMICAL   | NOAEL (WB)         | LOAEL (WB)      | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES      | LIFESTAGE         | EXPOSURE MODE | EXPOSURE DURATION | ENDPOINT         | ENDPOINT EFFECT                         | SOURCE                    | NOTES  |
|--|--------------------|-----------------|--------------|-------------|-------------|-----------|-------------------|-------------------|---------------|-------------------|------------------|---|---------------------------|--|
| gamma-Hexachloro-cyclohexane (lindane)   | 1,580 <sup>c</sup> | 79,000          |              |             |             | µg/kg     | sheepshead minnow | 17-21 mm          |               |                   | survival         | LR50                                    | Schimmel et al. (1977b)   | LR50   |
| Heptachlor (tech heptachlor: 65% heptachlor; 22% trans-chlordane; 2% cis-chlordane; 2% nonachlor)  | 150 <sup>b</sup>   | 1,500           |              |             |             | µg/kg     | spot              |                   | water         | 96 hrs            | survival         | LOAEL- 25% mortality                    | Schimmel et al. (1976)    | no statistical methods; saltwater study  |
| Heptachlor (technical heptachlor)  |                    | 1,900 - 2,600   |              |             |             | µg/kg     | bluegill          |                   | food          | up to 140 days    | growth           | average body weight                     | Andrews et al. (1966)     | fish were exposed to 5 ppm technical heptachlor; growth effects were not statistically evaluated; LOAEL is range of heptachlor residue concentrations (in two replicates) where average body weight was reduced (approx 90% of controls) after 28 days; growth e   |
| Heptachlor   | 1,700              | 5,300           |              |             |             | µg/kg     | spot              |                   | water         | 96 hrs            | survival         | NOAEL-0% mortality; LOAEL 85% mortality | Schimmel et al. (1976)    | no statistical methods; saltwater study  |
| Heptachlor (technical heptachlor)  |                    | 17,300 - 24,800 |              |             |             | µg/kg     |                   |                   | water         | 24 hrs            | survival         | 90% mortality                           | Andrews et al. (1966)     | fish were exposed to technical heptachlor; 90% mortality was observed after 72 hrs following exposure period in fish exposed to 0.0500 and 0.0375 ppm technical heptachlor; LOAEL is tissue residues at these dose levels at 72 hrs; heptachlor epoxide, other related compounds and DDT were also measured in fish tissue   |
| Heptachlor   |                    | 20,000          |              |             |             | µg/kg     | sheepshead minnow |                   | water         | 96 hrs            | survival         | LOAEL- 35% mortality                    | Schimmel et al. (1976)    | no statistical methods; saltwater study  |
| Heptachlor (tech heptachlor: 65% heptachlor; 22% trans-chlordane; 2% cis-chlordane; 2% nonachlor)\ | 5,700              | 34,000          |              |             |             | µg/kg     | pinfish           |                   | water         | 96 hrs            | survival         | NOAEL-5% mortality; LOAEL-50% mortality | Schimmel et al. (1976)    | no statistical methods; saltwater study  |
| Heptachlor epoxide (tech heptachlor)   | 80 <sup>b</sup>    | 800 - 900       |              |             |             | µg/kg     | bluegill          |                   | food          | up to 140 days    | growth           | average body weight                     | Andrews et al. (1966)     | fish were exposed to 5 ppm technical heptachlor; growth effects were not statistically evaluated; LOAEL is range of heptachlor residue concentrations (in two replicates) where average body weight was reduced (approx 90% of controls) after 28 days; growth effects were more pronounced at higher dose levels and after greater exposure duration; heptachlor, other related compounds and DDT were also measured in fish tissue |
| Heptachlor epoxide (tech heptachlor)   |                    | 1,660 - 2,400   |              |             |             | µg/kg     | bluegill          |                   | water         | 24 hrs            | survival         | 90% mortality                           | Andrews et al. (1966)     | fish were exposed to technical heptachlor; 90% mortality was observed after 72 hrs following exposure period in fish exposed to 50 and 37.5 µg/kg technical heptachlor; LOAEL is tissue residues at these dose levels at 72 hrs; heptachlor, other related compounds and DDT were also measured in fish tissue   |
| Hexachlorobenzene  | 46,500             |                 |              |             |             | µg/kg     | fathead minnow    |                   | water         | 28-day exposure   | survival         |   | Nebeker et al. (1989)     |  |
| Hexachlorobenzene  | 46,500             |                 |              |             |             | µg/kg     | fathead minnow    |                   | water         | 28-day exposure   | growth           |   | Nebeker et al. (1989)     |  |
| Hexachlorobenzene  | 97,000             |                 |              |             |             | µg/kg     | fathead minnow    | embryo - juvenile | water         | 32 days           | growth, survival | final body weight, survival             | Carlson and Kosian (1987) | no significant effect on growth or survival  |
| Hexachlorobenzene  | 97,000             |                 |              |             |             | µg/kg     | fathead minnow    | embryo            | water         | 32-day exposure   | growth, survival |   | Carlson and Kosian (1987) | development/mortality; embryos 4-12 hrs old; exposure period during embryo-to-juvenile stage; NOAEL is mean tissue residue n=2   |

Table 2, continued

| CHEMICAL                     | NOAEL (WB)         | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES            | LIFESTAGE | EXPOSURE MODE      | EXPOSURE DURATION   | ENDPOINT                        | ENDPOINT EFFECT  | SOURCE                       | NOTES   |
|------------------------------|--------------------|------------|--------------|-------------|-------------|-----------|-------------------------|-----------|--------------------|---|---------------------------------|--|------------------------------|---|
| Hexachlorobenzene            | 282,000            |            |              |             |             | µg/kg     | fathead minnow          |           | sediment and water | 28-day exposure   | survival                        |  | Schuytema et al. (1990)      | exposure period followed by 15-28 day depuration; uptake/bioaccumulation study  |
| Hexachlorobenzene            | 468,000            |            |              |             |             | µg/kg     | fathead minnow          |           | water              | 28-day exposure   | survival                        |  | Schuytema et al. (1990)      | exposure period followed by 15-28 day depuration; uptake/bioaccumulation study  |
| Methoxychlor                 | 50                 | 300        |              |             |             | µg/kg     | brook trout             | yearlings | food               | 30 days   | growth                          | body weight  | Oladimeji and Leduc (1975)   | fish were fed four different concentrations of methoxychlor at four different rates resulting in the same dose of 0.67 mg/kg bw/d; LOAEL is WB concentration in fish fed at 1% ingestion rate   |
| 4-methylphenol               | 1,530 <sup>c</sup> | 76,500     |              |             |             | µg/kg     | rainbow trout           |           | injection          | 96 hrs  | survival                        | LD50   | Kaiser et al. (1984)         | LOAEL is injected LD50 dose following 96-hr exposure; tissue residues not measured  |
| 4-methylphenol               |                    | 78,900     |              |             |             | µg/kg     | rainbow trout           |           | injection          | 96 hrs  | survival                        | LD50   | Hodson et al. (1988)         | LOAEL is injected LD50 dose following 96-hr exposure  |
| Phenol                       | 1,470 <sup>c</sup> | 73,400     |              |             |             | µg/kg     | rainbow trout           |           | water              | 13 hrs  | survival                        | lethal body burden   | McKim & Schmieder (1990)     | LOAEL is reported lethal residue - limited detail on study methods  |
| Phenol                       | 76,000             | 114,000    |              |             |             | µg/kg     | goldfish                |           | water              | 25 hrs  | survival                        | LC50   | Kobayashi et al. (1979)      | LOAEL is LC50 body burden based on dead fish exposed to 60 µg/L; NOAEL is estimated body burden using BCF reported in study (1.9) and water concentration (40 µg/L) where 95% survival was reported over the duration of the experiment |
| Phenol                       |                    | 238,000    |              |             |             | µg/kg     | goldfish                |           | water              | 5 hrs   | survival                        | LC50   | Kishino and Kobayashi (1995) | LOAEL is LC 50 from pH 8 study; LC50s were 273, 238, and 420 µg/g at pH 6, 8 and 10, respectively; mean residues of dead fish from all exposure levels from pH levels 6-10 were: 223 - 309 µg/g   |
| Phenol                       |                    | 420,000    |              |             |             | µg/kg     | rainbow trout           |           | injection          | 96 hrs  | survival                        | LD50   | Kaiser et al. (1984)         | LOAEL is injected LD50 dose following 96-hr exposure; tissue residues not measured  |
| Bis (2-ethylhexyl) phthalate | 390                |            |              |             |             | µg/kg     | rainbow trout - sac fry |           |                    | 10 day prior to hatch and effects measured 24 days post hatch | reproduction (sac-fry survival) | 13% mortality at NOAEL (not statistically different than control); 6% mortality in control | Mehrle and Mayer (1976)      | fry tissue residues calculated from bioconcentration factor and water concentrations reported in paper.   |
| Bis (2-ethylhexyl) phthalate | 1,000 - 6,750      |            |              |             |             | µg/kg     | sheepshead minnow       |           | water              | 24 hrs  | survival                        | survival   | Wofford et al. (1981)        | NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported  |
| Butylbenzyl phthalate        | 6,450              |            |              |             |             | µg/kg     | bluegill                |           | water              | 21 days   | survival                        | survival   | Barrows et al. (1980)        | NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported; study is a bioaccumulation study  |
| Di(n)butyl phthalate         | 660                |            |              |             |             | µg/kg     | bluegill                |           | water              | 42 days   | survival                        | survival   | Barrows et al. (1980)        | NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported; study is a bioaccumulation study  |
| Di(n)butyl phthalate         | 1,170              |            |              |             |             | µg/kg     | sheepshead minnow       |           | water              | 24 hrs  | survival                        | survival   | Wofford et al. (1981)        | NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported  |
| Dimethyl phthalate           | 498                |            |              |             |             | µg/kg     | bluegill                |           | water              | 21 days   | survival                        | survival   | Barrows et al. (1980)        | NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported; study is a bioaccumulation study  |
| Diethyl phthalate            | 1,102              |            |              |             |             | µg/kg     | bluegill                |           | water              | 21 days   | survival                        | survival   | Barrows et al. (1980)        | NOAEL is based on assumed 100% survival; NOAEL is based on the assumption that no mortality was observed because none was reported; study is a bioaccumulation study  |

Table 2, continued

| CHEMICAL     | NOAEL (WB) | LOAEL (WB) | CONV. FACTOR | NOAEL (egg) | LOAEL (egg) | UNIT (ww) | TEST SPECIES  | LIFESTAGE | EXPOSURE MODE | EXPOSURE DURATION | ENDPOINT | ENDPOINT EFFECT | SOURCE                | NOTES  |
|--------------|------------|------------|--------------|-------------|-------------|-----------|---------------|-----------|---------------|-------------------|----------|-----------------|-----------------------|--|
| Benzoic acid | 3,380      |            |              |             |             | µg/kg     | mosquito fish |           | water         | 24 hrs            | survival | survival        | Lu and Metcalf (1975) | highly uncertain study- exposure conditions was a mini-ecosystem with no mortality results discussed (bioaccumulation study); was assumed that no mortality was observed because none was reported |

<sup>a</sup> NOAEL estimated using uncertainty factor of 5 (chronic LOAEL to chronic NOAEL).

<sup>b</sup> NOAEL estimated using uncertainty factor of 10 (acute LOAEL to chronic NOAEL).

<sup>c</sup> NOAEL estimated using uncertainty factor of 50 (LC50 to chronic NOAEL).

LC10 – concentration that causes the death of 10% of a group of test animals

LC50 – concentration that causes the death of 50% of a group of test animals

LOAEL – lowest-observed-adverse-effect level

NC – TRVs not reported in database because study only injection dose was reported (no WB tissue residues were reported)

NOAEL – no-observed-adverse-effect level

PCB – polychlorinated biphenyl

TCDD – tetrachlorodibenzo-*p*-dioxin

TBT – tributyltin

ww – concentration in fish tissue was reported on a wet weight basis

## REFERENCES

- Adams WJ, DeGraeve GM, Sabourin TD, Cooney JD, Mosher GM. 1986. Toxicity and bioconcentration of 2,3,7,8-TCDD to fathead minnows (*Pimephales promelas*). *Chemosphere* 15(9-12):1503-1511.
- Allison D, Kallman BJ, Cope OB, Van Valin C. 1963. Insecticides: effects on cutthroat trout of repeated exposure to DDT. *Science* 142(3594):958-961.
- Allison DT, Kollman BJ, Cope OB, Van Valin C. 1964. Some chronic effects of DDT on cutthroat trout. Research report 64. Bureau of Sport Fisheries and Wildlife, US Fish and Wildlife Service, Washington, DC.
- Andrews AK, Van Valin CC, Stebbings BE. 1966. Some effects of heptachlor on bluegills. *Trans Am Fish Soc* 95:297-309.
- Argyle RL, Williams GC, Dupree HK. 1973. Endrin uptake and release by fingerling channel catfish (*Ictalurus punctatus*). *J Fish Res Board Can* 30:1743-1744.
- Baker RTM, Handy RD, Davies SJ, Snook JC. 1998. Chronic dietary exposure to copper affects growth, tissue lipid peroxidation, and metal composition of the grey mullet, *Chelon labrosus*. *Mar Environ Res* 45(4/5):357-365.
- Baldisserotto B, Chowdhury MJ, Wood CM. 2005. Effects of dietary calcium and cadmium on cadmium accumulation, calcium and cadmium uptake from the water, and their interactions in juvenile rainbow trout. *Aquat Toxicol* 72:99-117.
- Barrows ME, Petrocelli SR, Macek KJ, Carroll JR. 1980. Bioconcentration and elimination of selected water pollutants by bluegill sunfish (*Lepomis macrochirus*). In: Haque R, ed, Dynamics, exposure and hazard assessment of toxic chemicals. Ann Arbor Science Publishers, Ann Arbor, MI.
- Benton MJ, Nimrod AC, Benson WH. 1994. Evaluation of growth and energy storage as biological markers of DDT exposure in sailfin mollies. *Ecotoxicol Environ Saf* 29:1-12.
- Berntssen MHG, Lundebye AK, Maage A. 1999a. Effects of elevated dietary copper concentrations on growth, feed utilization and nutritional status of Atlantic salmon (*Salmo salar* L.) fry. *Aquaculture* 174:167-181.
- Berntssen MHG, Hylland K, Bonga SEW, Maage A. 1999b. Toxic levels of dietary copper in Atlantic salmon (*Salmo salar* L.) parr. *Aquat Toxicol* 46:87-99.
- Birge WJ, Black JA, Westerman AG, Hudson JE. 1979. The effects of mercury on reproduction of fish and amphibians. In: Nriagu JO, ed, The biogeochemistry of mercury in the environment. Elsevier/North-Holland Biomedical Press, New York, NY, pp 629-655.
- Blazer VS, Fournie JW, Weeks-Perkins BA. 1997. Macrophage aggregates: biomarker for immune function in fishes? In: Dwyer FJ, Doane TR, Hinman ML, eds,



Environmental toxicology and risk assessment: Modeling and risk assessment. Vol 6. STP 1317. American Society for Testing and Materials, Philadelphia, PA, pp 360-375.

Branson DR, Takahashi IT, Parker WM, Blau GE. 1985. Bioconcentration kinetics of 2,3,7,8-tetrachlorodibenzo-p-dioxin in rainbow trout. *Environ Toxicol Chem* 4:779-788.

Buhler DR, Rasmusson ME, Shanks WE. 1969. Chronic oral DDT toxicity in juvenile coho and chinook salmon. *Toxicol Appl Pharmacol* 14:535-555.

Butler PA. 1969. Significance of DDT residues in estuarine fauna. In: Miller MW, Berg GG, eds, *Chemical fallout: current research on persistent pesticides*. Charles C. Thomas, Springfield, IL, pp 205-220.

Carlson AR, Kosian PA. 1987. Toxicity of chlorinated benzenes to fathead minnows (*Pimephales promelas*). *Arch Environ Contam Toxicol* 16:129-135.

Cockell KA, Bettger WJ. 1993. Investigations of the gallbladder pathology associated with dietary exposure to disodium arsenate heptahydrate in juvenile rainbow trout. *Toxicology* 77:233-248.

Cockell KA, Hilton JW. 1988. Preliminary investigation on the comparative chronic toxicity of four dietary arsenicals to juvenile rainbow trout (*Salmo gairdneri* R.). *Aquat Toxicol* 12:73-82.

Cockell KA, Hilton JW, Bettger WJ. 1991. Chronic toxicity of dietary disodium arsenate heptahydrate to juvenile rainbow trout (*Oncorhynchus mykiss*). *Arch Environ Contam Toxicol* 21:518-527.

Cockell KA, Hilton JW, Bettger WJ. 1992. Hepatobiliary and hematological effects of dietary disodium arsenate heptahydrate in juvenile rainbow trout (*Oncorhynchus mykiss*). *Comp Biochem Physiol* 103C(3):453-458.

Courtney CH, Reed JK. 1971. Accumulation of DDT from food and from water by golden shiner minnows, *Notemigonus crysoleucas*. 25th Annual Conference, Southeastern Association of Game and Fish Commissioners, Charleston SC. Southeastern Association of Game and Fish Commissioners, Frankfort, KY, pp 426-431.

Crawford RB, Guarino AM. 1976. Effects of DDT in *Fundulus*: studies on toxicity, fate, and reproduction. *Arch Environ Contam Toxicol* 4:334-348.

Elonen GE, Spehar RL, Holcombe GW, Johnson RD, Fernandez JD, Erickson RJ, Tietge JE, Cook PM. 1998. Comparative toxicity of 2,3,7,8-tetrachlorodibenzo-p-dioxin to seven freshwater fish species during early life-stage development. *Environ Toxicol Chem* 17(3):472-483.

EPA. 2004. Draft aquatic life water quality criteria for selenium - 2004. EPA-822-D-04-001. US Environmental Protection Agency, Washington, DC.

- Fabacher DL. 1976. Toxicity of endrin and an endrin-methyl parathion formulation to largemouth bass fingerlings. *Bull Environ Contam Toxicol* 16(3):376-378.
- Feroz M, Khan MAQ. 1979. Fate of  $^{14}\text{C}$ -cis-chlordane in goldfish, *Carassius auratus* (L.). *Bull Environ Contam Toxicol* 23:64-69.
- Fisher JP, Spitsbergen JM, Bush B, Jahan-Parwar B. 1994. Effect of embryonic PCB exposure on hatching success, survival, growth and developmental behavior in landlocked Atlantic salmon, *Salmo salar*. In: Gorsuch JW, Dwyer FJ, Ingersoll CG, La Point TW, eds, *Environmental toxicology and risk assessment*. Vol 2. ASTM STP 1216. American Society for Testing and Materials, Philadelphia, PA, pp 298-314.
- Fisk AT, Yarechewski AL, Metner DA, Evans RE, Lockhart WL, Muir DCG. 1997. Accumulation, depuration and hepatic mixed-function oxidase enzyme induction in juvenile rainbow trout and lake whitefish exposed to dietary 2,3,7,8-tetrachlorodibenzo-p-dioxin. *Aquat Toxicol* 37:201-220.
- Fjeld E, Haugen TO, Vollestad LA. 1998. Permanent impairment in the feeding behavior of grayling (*Thymallus thymallus*) exposed to methylmercury during embryogenesis. *Sci Total Environ* 213:247-254.
- Franklin NM, Glover CN, Nicol JA, Wood CM. 2005. Calcium/cadmium interactions at uptake surfaces in rainbow trout: waterborne versus dietary routes of exposure. *Environ Toxicol Chem* 24(11):2954-2964.
- Freeman HC, Idler DR. 1975. The effect of PCB on steroidogenesis and reproduction in the brook trout (*Salvelinus fontinalis*). *Can J Biochem* 53:666-670.
- Galvez F, Wood CM. 1999. Physiological effects of dietary silver sulfide exposure in rainbow trout. *Environ Toxicol Chem* 18(1):84-88.
- Gatlin DM, III, Wilson RP. 1986. Dietary copper requirement of fingerling channel catfish. *Aquacult* 54:277-285.
- Giesy JP, Jones PD, Kannan K, Newsted JL, Tillitt DE, Williams LL. 2002. Effects of chronic dietary exposure to environmentally relevant concentrations of 2,3,7,8-tetrachlorodibenzo-p-dioxin on survival, growth, reproduction and biochemical responses of female rainbow trout (*Oncorhynchus mykiss*). *Aquat Toxicol* 59:35-53.
- Goettl JP, Davies PH, Sinley JR. 1976. Water pollution studies. Colorado Fisheries Research Review 1972-1975. *CO Div Wildlife Rev* 8:68-75.
- Guiney PD, Cook PM, Casselman JM, Fitzsimmons JD, Simonin HA, Zabel EW, Peterson RE. 1996. Assessment of 2,3,7,8-tetrachlorodibenzo-p-dioxin induced sac fry mortality in lake trout (*Salvelinus namaycush*) from different regions of the Great Lakes. *Can J Fish Aquat Sci* 53:2080-2092.

- Hamelink JL, Waybrant RC, Ball RC. 1971. A proposal: exchange equilibria control the degree chlorinated hydrocarbons are biologically magnified in lentic environments. *Trans Am Fish Soc* 100(2):207-213.
- Handy RD. 1992. The assessment of episodic metal pollution. II. The effects of cadmium and copper enriched diets on tissue contaminant analysis in rainbow trout (*Oncorhynchus mykiss*). *Arch Environ Contam Toxicol* 22:82-87.
- Handy RD. 1993. The effect of acute exposure to dietary Cd and Cu on organ toxicant concentrations in rainbow trout, *Oncorhynchus mykiss*. *Aquat Toxicol* 27:1-14.
- Hansen D, Schimmel SC, Forester J. 1974a. Aroclor 1254 in eggs of sheepshead minnows: effect on fertilization success and survival of embryos and fry. *Proceedings of 27th Annual Conference, Southeastern Association of Game and Fish Commissioners, Hot Springs, AR*, pp 420-426.
- Hansen DJ, Parrish PR, Lowe JL, Wilson AJ, Jr, Wilson PD. 1971. Chronic toxicity, uptake, and retention of Aroclor 1254 in two estuarine fishes. *Bull Environ Contam Toxicol* 6:113-119.
- Hansen DJ, Parrish PR, Forester J. 1974b. Aroclor 1016: Toxicity to and uptake by estuarine animals. *Environ Res* 7:363-373.
- Hansen DJ, Schimmel SC, Forester J. 1975. Effects of Aroclor 1016 on embryos, fry, juveniles, and adults of sheepshead minnows (*Cyprinodon variegates*). *Trans Am Fish Soc* 104:584-588.
- Hansen DJ, Schimmel SC, Forester J. 1977. Endrin: effects on the entire life cycle of a saltwater fish, *Cyprinodon variegatus*. *J Toxicol Environ Health* 3:721-733.
- Hart DR, Heddle JA. 1991. Micronucleus assays in peripheral blood of rainbow trout: Timing of response and chemical mutagen sensitivity. *Can Tech Rep Fish Aquat Sci* 1774(2):993-1010.
- Hatakeyama S, Yasuno M. 1982. Accumulation and effects of cadmium on guppy (*Poecilia reticulata*) fed cadmium-dosed Cladocera (*Moina macrocopa*). *Bull Environ Contam Toxicol* 29:159-166.
- Hatakeyama S, Yasuno M. 1987. Chronic effects of Cd on the reproduction of the guppy (*Poecilia reticulata*) through Cd-accumulated midge larvae (*Chironomus yoshimatsui*). *Ecotoxicol Environ Saf* 14:191-207.
- Hattula ML, Karlog O. 1972. Toxicity of polychlorinated biphenyls (PCBs) to goldfish. *Acta Pharmacol Toxicol* 31:238-240.
- Hawkes CL, Norris LA. 1977. Chronic oral toxicity of 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) to rainbow trout. *Trans Am Fish Soc* 106(6):641-645.
- Heisinger JF, Green W. 1975. Mercuric chloride uptake by eggs of the ricefish and resulting teratogenic effects. *Bull Environ Contam Toxicol* 14(6):665-673.

- Heisinger JF, Hansen CD, Kim JH. 1979. Effect of selenium dioxide on the accumulation and acute toxicity of mercuric chloride in goldfish. *Arch Environ Contam Toxicol* 8:279-283.
- Hendricks JD, Scott WT, Putnam TP, Sinnhuber RO. 1981. Enhancement of aflatoxin B1 hepatocarcinogenesis in rainbow trout (*Salmo gairdneri*) embryos by prior exposure of gravid females to dietary Aroclor 1254. In: Branson DR, Dickson KL, eds, *Aquat Toxicol and Hazard Assessment. Fourth Conference. ASTM STP 737*. American Society for Testing and Materials, Philadelphia, PA, pp 203-214.
- Hendricks JD, Meyers TR, Shelton DW, Casteel JL, Bailey GS. 1985. Hepatocarcinogenicity of benzo[a]pyrene to rainbow trout by dietary exposure and intraperitoneal injection. *J Nat Cancer Inst* 74:839-851.
- Hilton JW, Bettger WJ. 1988. Dietary vanadium toxicity in juvenile rainbow trout: a preliminary study. *Aquat Toxicol* 12:63-71.
- Hodson PV, Dixon DG, Kaiser KLE. 1988. Estimating the acute toxicity of waterborne chemicals in trout from measurements of median lethal dose and the octanol-water partition coefficient. *Environ Toxicol Chem* 7:443-454.
- Hugla JL, Thome JP. 1999. Effects of polychlorinated biphenyls on liver ultrastructure, hepatic monooxygenases, and reproductive success in the barbel. *Ecotoxicol Environ Saf* 42:265-273.
- Jarvinen AW, Tyo RM. 1978. Toxicity to fathead minnows of endrin in food and water. *Arch Environ Contam Toxicol* 7:409-421.
- Jarvinen AW, Hoffman MJ, Thorslund TW. 1976. Toxicity of DDT food and water exposure to fathead minnows. EPA-600/3-76/114. US Environmental Protection Agency, Duluth, MN.
- Jarvinen AW, Hoffman MJ, Thorslund TW. 1977. Long-term toxic effects of DDT food and water exposure on fathead minnows (*Pimephales promelas*). *J Fish Res Board Can* 34:2089-2103.
- Johnson RD, Tietge JE, Jensen KM, Fernandez JD, Linnum AL, Lothenback DB, Holcombe GW, Cook PM, Christ SA, Lattier DL, Gordon DA. 1998. Toxicity of 2,3,7,8-tetrachlorodibenzo-*p*-dioxin to early life stage brook trout (*Salvelinus fontinalis*) following parental dietary exposure. *Environ Toxicol Chem* 17(12):2408-2421.
- Kaiser KLE, Dixon DG, Hodson PV. 1984. QSAR studies on chlorophenols, chlorobenzenes and para-substituted phenols. In: Kaiser KLE, ed, *QSAR in environmental toxicology*. D. Reidel Publishing Co., Dordrecht, the Netherlands, pp 189-206.
- Kamunde CN, Grosell M, Lott JNA, Wood CM. 2001. Copper metabolism and gut morphology in rainbow trout (*Oncorhynchus mykiss*) during chronic sublethal dietary copper exposure. *Can J Fish Aquat Sci* 58:293-305.

- Kang J-C, Kim S-G, Jang S-W. 2005. Growth and hematological changes of rockfish, *Sebastes schlegeli* (Hilgendorf) exposed to dietary Cu and Cd. J World Aquacult Soc 36(2):188-195.
- Kim S-G, Kim J-W, Kang J-C. 2004. Effect of dietary cadmium on growth and haematological parameters of juvenile rockfish, *Sebastes schlegeli* (Hilgendorf). Aquacult Res 35:80-86.
- Kishino T, Kobayashi K. 1995. Relation between toxicity and accumulation of chlorophenols at various pH, and their absorption mechanism in fish. Wat Res 29(2):431-442.
- Kobayashi K, Akitake H, Manabe K. 1979. Relation between toxicity and accumulation of various chlorophenols in goldfish. Bull Jap Soc Sci Fish 45(2):173-175.
- Kudo A, Mortimer DC. 1979. Pathways for mercury uptake by fish from bed sediments. Environ Pollut 13(1979):239-245.
- Lanno RP, Slinger SJ, Hilton JW. 1985a. Effect of ascorbic acid on dietary copper toxicity in rainbow trout (*Salmo gairdneri* Richardson). Aquaculture 49:269-287.
- Lanno RP, Slinger SJ, Hilton JW. 1985b. Maximum tolerable and toxicity levels of dietary copper in rainbow trout (*Salmo gairdneri* Richardson). Aquaculture 49:257-268.
- Latif MA, Bodaly RA, Johnston TA, Fudge RJP. 2001. Effects of environmental and maternally derived methylmercury on the embryonic and larval stages of walleye (*Stizostedion vitreum*). Environ Pollut 111(2001):139-148.
- Lieb AJ, Bills DD, Sinnhuber RO. 1974. Accumulation of dietary polychlorinated biphenyls (Aroclor 1254) by rainbow trout (*Salmo gairdneri*). J Agr Food Chem 22:638-642.
- Lock RAC. 1975. Uptake of methylmercury by aquatic organisms from water and food. In: Koeman JH, Strik JJTWA, eds, Sublethal effects of toxic chemicals on aquatic organisms. Elsevier Press, Amsterdam, pp 61-79.
- Lorentzen M, Maage A, Julshamn K. 1998. Supplementing copper to a fish meal based diet fed to Atlantic salmon parr affects liver copper and selenium concentrations. Aquacult Nutrit 4:67-72.
- Lu P-Y, Metcalf RL. 1975. Environmental fate and biodegradability of benzene derivatives as studied in a model aquatic ecosystem. Environ Health Persp 10:268-284.
- Ludke JL, Ferguson DE, Burke WD. 1968. Some endrin relationships in resistant and susceptible populations of golden shiners, *Notemigonus crysoleucas*. Trans Am Fish Soc 97:260-263.

- Lundebye AK, Berntssen MHG, Bonga WSE, Maage A. 1999. Biochemical and physiological responses in Atlantic salmon (*Salmo salar*) following dietary exposure to copper and cadmium. *Mar Pollut Bull* 39:137-144.
- Macek KJ. 1968a. Growth and resistance to stress in brook trout fed sublethal levels of DDT. *J Fish Res Bd Can* 25(11):2443-2451.
- Macek KJ. 1968b. Reproduction in brook trout (*Salvelinus fontinalis*) fed sublethal concentrations of DDT. *J Fish Res Board Can* 25(9):1787-1796.
- Macek KJ, Korn S. 1970. Significance of the food chain in DDT accumulation by fish. *J Fisheries Res Board Can* 27(8):1496-1498.
- Macek KJ, Rodgers CR, Stalling DL, Korn S. 1970. The uptake, distribution and elimination of dietary  $^{14}\text{C}$ -DDT and  $^{14}\text{C}$ -Dieldrin in rainbow trout. *Trans Am Fish Soc* 99:689-695.
- Matta MB, Linse J, Cairncross C, Francendese L, Kocan RM. 2001. Reproductive and transgenerational effects of methylmercury or Aroclor 1268 on *Fundulus heteroclitus*. *Environ Toxicol Chem* 20(2):327-335.
- Mauck WL, Mehrle PM, Mayer FL. 1978. Effects of the polychlorinated biphenyl Aroclor 1254 on growth, survival, and bone development in brook trout (*Salvelinus fontinalis*). *J Fish Res Board Can* 35:1084-1088.
- Mayer KS, Mayer FL, Witt A, Jr. 1985. Waste transformer oil and PCB toxicity to rainbow trout. *Trans Am Fish Soc* 114(6):869-886.
- McCarthy ID, Fuiman LA, Alvarez MC. 2003. Aroclor 1254 affects growth and survival skills of Atlantic croaker *Micropogonias undulatus* larvae. *Mar Ecol Prog Ser* 252:295-301.
- McKim JM, Schneider PK. 1990. Bioaccumulation: does it reflect toxicity? In: Nagel R, Loskill R, eds, *Bioaccumulation in aquatic systems: Contributions to the assessment*. Proceedings, International Workshop, Berlin, 1990. VCG Publishers, New York, NY, pp 161-188.
- McKim JM, Olson GF, Holcombe GW, Hunt EP. 1976. Long-term effects of methylmercuric chloride on three generations of brook trout (*Salvelinus fontinalis*): Toxicity, accumulation, distribution, and elimination. *J Fish Res Board Can* 33:2726-2739.
- Meador JP, Sommers FC, Ylitalo GM, Sloan CA. 2006. Altered growth and related physiological responses in juvenile chinook salmon (*Oncorhynchus tshawytscha*) from dietary exposure to polycyclic aromatic hydrocarbons (PAHs). *Can J Fish Aquat Sci* 63:2364-2376.
- Mehrle PM, Mayer FL. 1976. Di-2-ethylhexyl phthalate: residue dynamics and biological effects in rainbow trout and fathead minnows. *Proceedings of*

University of Missouri's 10th Annual Conference on Trace Substances in Environmental Health, June 8-10, Columbia, MO, pp 519-524.

- Mehrle PM, Buckler DR, Little EE, Smith LM, Petty JD, Peterman PH, Stalling DL, De Graeve GM, Coyle JJ, Adams WJ. 1988. Toxicity and bioconcentration of 2,3,7,8-tetrachlorodibenzodioxin and 2,3,7,8-tetrachlorodibenzofuran in rainbow trout. *Environ Toxicol Chem* 7:47-62.
- Miller PA, Lanno RP, McMaster ME, Dixon DG. 1993. Relative contributions of dietary and waterborne copper to tissue copper burdens and waterborne-copper tolerance in rainbow trout (*Oncorhynchus mykiss*). *Can J Fish Aquat Sci* 5(8):1683-1689.
- Miller RA, Norris LA, Loper BR. 1979. The response of coho salmon and guppies to 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) in water. *Trans Am Fish Soc* 108:401-407.
- Moore R, Toro E, Stanton M, Khan MAQ. 1977. Absorption and elimination of <sup>14</sup>C-*alpha*- and *gamma*-chlordanes by a freshwater alga, daphnid, and goldfish. *Arch Environ Contam Toxicol* 6:411-420.
- Mount DR, Barth AK, Garrison TD, Barten KA, Hockett JR. 1994. Dietary and waterborne exposure of rainbow trout (*Oncorhynchus mykiss*) to copper, cadmium, lead and zinc using a live diet. *Environ Toxicol Chem* 13(12):2031-41.
- Muller JK, Johnson KG, Sepulveda MS, Borgert CJ, Gross TS. 2004. Accumulation of dietary DDE and dieldrin by largemouth bass, *Micropterus salmoides floridanus*. *Bull Environ Contam Toxicol* 73:1078-1085.
- Murai T, Andrews JW, Smith RG, Jr. 1981. Effects of dietary copper on channel catfish. *Aquaculture* 22:353-357.
- Nakayama K, Oshima Y, Nagafuchi K, Hano T, Shimasaki Y, Honjo T. 2005. Early-life-stage toxicity in offspring from exposed parent medaka, *Oryzias latipes*, to mixtures of tributyltin and polychlorinated biphenyls. *Environ Toxicol Chem* 24(3):591-596.
- Nebeker AV, Puglisi FA, DeFoe DL. 1974. Effect of polychlorinated biphenyl compounds on survival and reproduction of the fathead minnow and flagfish. *Trans Am Fish Soc* 103:562-568.
- Nebeker AV, Schuytema GS, Griffis WL, Barbitta JA, Carey LA. 1989. Effect of sediment organic carbon on survival of *Hyaella azteca* exposed to DDT and endrin. *Environ Toxicol Chem* 8:705-18.
- Niimi AJ, Lowe-Jinde L. 1984. Differential blood cell ratios of rainbow trout (*Salmo gairdneri*) exposed to methylmercury and chlorobenzenes. *Arch Environ Contam Toxicol* 13:303-311.

- Nirmala K, Oshima Y, Lee R, Imada N, Honjo T, Kobayashi K. 1999. Transgenerational toxicity of tributyltin and its combined effects with polychlorinated biphenyls on reproductive processes in Japanese medaka (*Oryzias latipes*). *Environ Toxicol Chem* 18(4):717-721.
- Oladimeji AA, Leduc G. 1975. Effects of dietary methoxychlor on the food maintenance requirements of brook trout. *Prog Wat Tech* 7(3/4):587-598.
- Oladimeji AA, Qadri SU, DeFreitas ASW. 1984. Long-term effects of arsenic accumulation in rainbow trout, *Salmo gairdneri*. *Bull Environ Contam Toxicol* 32:732-741.
- Olson GF, Mount DI, Snarski VM, Thorslund TW. 1975. Mercury residues in fathead minnows, *Pimephales promelas* Rafinesque, chronically exposed to methylmercury in water. *Bull Environ Contam Toxicol* 14(2):129-134.
- Palm RC, Jr, Powell DG, Skillman A, Godtfredsen K. 2003. Immunocompetence of juvenile chinook salmon against *Listonella anguillarum* following dietary exposure to polycyclic aromatic hydrocarbons. *Environ Toxicol Chem* 22(12):2986-2994.
- Parrish PR, Schimmel SC, Hansen DJ, Patrick Jr JM, Forester J. 1976. Chlordane: effects on several estuarine organisms. *J Toxicol Environ Health* 1:485-494.
- Phillips GR, Buhler DR. 1978. The relative contributions of methylmercury from food or water to rainbow trout (*Salmo gairdneri*) in a controlled laboratory environment. *Trans Am Fish Soc* 107(6):853-861.
- Pillai MKK, Agarwal HC, Yadav DV. 1977. Tolerance, uptake and metabolism of DDT in *Gambusia affinis*. *Ind J Exper Biol* 15:40-41.
- Powell DB, Palm RC, Jr, Skillman A, Godtfredsen K. 2003. Immunocompetence of juvenile chinook salmon against *Listonella anguillarum* following dietary exposure to Aroclor 1254. *Environ Toxicol Chem* 22(2):285-295.
- Prince R, Cooper KR. 1995. Comparisons of the effects of 2,3,7,8-tetrachlorodibenzo-p-dioxin on chemically impacted and nonimpacted subpopulations of *Fundulus heracitus*: I. TCDD toxicity. *Environ Toxicol Chem* 14(4):579-587.
- Rhead MM, Perkins JM. 1984. An evaluation of the relative importance of food and water as sources of p,p'-DDT to the goldfish, *Carassius auratus* (L.). *Wat Res* 18(6):719-725.
- Rice CA, Myers MS, Willis ML, French BL, Casillas E. 2000. From sediment bioassay to fish biomarker – connecting the dots using simple trophic relationships. *Mar Environ Res* 50:527-533.
- Rodgers DW, Beamish FWH. 1982. Dynamics of dietary methylmercury in rainbow trout, *Salmo gairdneri*. *Aquat Toxicol* 2(1982):271-290.



- Schimmel SC, Patrick Jr JM, Forester J. 1976. Heptachlor: toxicity to and uptake by several estuarine organisms. *J Toxicol Environ Health* 1:955-965.
- Schimmel SC, Patrick Jr JM, Wilson Jr AJ. 1977a. Acute toxicity to and bioconcentration of endosulfan by estuarine animals. In: Mayer FL, Hamelink JL, eds, *Aquatic toxicology and hazard evaluation*, ASTM STP 634. American Society for Testing and Materials, Philadelphia, PA, pp 241-252.
- Schimmel SC, Patrick Jr JM, Forester J. 1977b. Toxicity and bioconcentration of BHC and lindane in selected estuarine animals. *Arch Environ Contam Toxicol* 6:355-363.
- Schmieder P, Lothenbach D, Tietge J, Erickson R, Johnson RF. 1995. [<sup>3</sup>H]-2,3,7,8-TCDD uptake and elimination kinetics of medaka (*Oryzias latipes*). *Environ Toxicol Chem* 14(10):1735-1743.
- Schuytema GS, Krawczyk DF, Griffis WL, Nebeker AV, Robideaux ML. 1990. Hexachlorobenzene uptake by fathead minnows and macroinvertebrates in recirculating sediment/water systems. *Arch Environ Contam Toxicol* 19:1-9.
- Shimasaki Y, Kitano T, Oshima Y, Inoue S, Imada N, Honjo T. 2003. Tributyltin causes masculinization in fish. *Environ Toxicol Chem* 22(1):141-144.
- Shubat PJ, Curtis LR. 1986. Ration and toxicant preexposure influence dieldrin accumulation by rainbow trout (*Salmo gairdneri*). *Environ Toxicol Chem* 5:69-77.
- Snarski VM, Olson GF. 1982. Chronic toxicity and bioaccumulation of mercuric chloride in the fathead minnow (*Pimephales promelas*). *Aquat Toxicol* 2:143-156.
- Spehar RL, Cook PM, Elonen GE, Holcombe GW, Johnson RD, Fernandez JD, Tietge JE. 1997. Comparison of early-life-stage sensitivity of freshwater fish exposed as eggs to 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD). Submitted to *Dioxin 97: 17th Symposium on Chlorinated Dioxins and Related Compounds*, August 25-29, 1997, Indianapolis, IN. US Environmental Protection Agency, Duluth, MN.
- Spitsbergen JM, Walker MK, Olson JR, Peterson RE. 1991. Pathologic alterations in early life stages of lake trout, *Salvelinus namaycush*, exposed to 2,3,7,8-tetrachlorodibenzo-p-dioxin as fertilized eggs. *Aquat Toxicol* 19:41-72.
- Szebedinsky C, McGeer JC, McDonald DG, Wood CM. 2001. Effects of chronic Cd exposure via the diet or water on internal organ-specific distribution and subsequent gill Cd uptake kinetics in juvenile rainbow trout (*Oncorhynchus mykiss*). *Environ Toxicol Chem* 20(3):597-607.
- Takeda H, Shimma Y. 1977. Effects of toxic amounts of dietary zinc on the growth and body components of rainbow trout at two levels of calcium. *Bull Freshw Fish Res* 27:103-109.
- Tietge JE, Johnson RD, Jensen KM, Cook PM, Elonen GE, Fernandez JD, Holcombe GW, Lothenbach DB, Nichols JW. 1998. Reproductive toxicity and disposition

- of 2,3,7,8-tetrachlorodibenzo-p-dioxin in adult brook trout (*Salvelinus fontinalis*) following a dietary exposure. Environ Toxicol Chem 17(12):2395-2407.
- van Wezel AP, Sijm DTHM, Seinen W, Opperhuizen A. 1995. Use of lethal body burdens to indicate species differences in susceptibility to narcotic toxicants. Chemosphere 31(5):3201-3209.
- Walker MK, Peterson RE. 1994. Toxicity of 2,3,7,8-tetrachlorodibenzo-p-dioxin to brook trout (*Salvelinus fontinalis*) during early development. Environ Toxicol Chem 13:817-820.
- Walker MK, Spitsbergen JM, Olson JR, Peterson RE. 1991. 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) toxicity during early life stage development of lake trout (*Salvelinus namaycush*). Can J Fish Aquat Sci 48:875-883.
- Walker MK, Hufnagle LC, Clayton MK, Peterson RE. 1992. An egg injection method for assessing early life stage mortality of polychlorinated dibenzo-p-dioxins, dibenzofurans, and biphenyls in rainbow trout (*Oncorhynchus mykiss*). Aquat Toxicol 22:15-38.
- Walker MK, Cook PM, Batterman AR, Butterworth BC, Berini C, Libal JJ, Hufnagle LC, Peterson RE. 1994. Translocation of 2,3,7,8-tetrachlorodibenzo-p-dioxin from adult female lake trout (*Salvelinus namaycush*) to oocytes: effects on early life stage development and sac fry survival. Can J Fish Aquat Sci 51:1410-1419.
- Walsh AR, O'Halloran J, Gower AM. 1994. Some effects of elevated levels of chromium (III) in sediments to the mullet *Chelon labrosus* (R). Ecotoxicol Environ Saf 27:168-176.
- Warlen SM, Wolfe DA, Lewis CW, Colby DR. 1977. Accumulation and retention of dietary <sup>14</sup>C-DDT by Atlantic menhaden. Trans Am Fish Soc 106(1):95-104.
- Webber HM, Haines TA. 2003. Mercury effects on predator avoidance behavior of a forage fish, golden shiner (*Notemigonus crysoleucas*). Environ Toxicol Chem 22(7):1556-1561.
- Wofford HW, Wilsey CD, Neff GS, Giam CS, Neff JM. 1981. Bioaccumulation and metabolism of phthalate esters by oysters, brown shrimp, and sheepshead minnows. Ecotoxicol Environ Saf 5:202-210.
- Wu RSS, Pollino CA, Au DWT, Zheng GJ, Yuen BBH, Lam PKS. 2003. Evaluation of biomarkers of exposure and effect in juvenile areolated grouper (*Epinephelus areolatus*) on foodborne exposure to benzo(a)pyrene. Environ Toxicol Chem 22(7):1568-1573.